

USING REMOTE SENSING TO CHARACTERIZE RIPARIAN VEGETATION: A REVIEW OF AVAILABLE TOOLS AND PERSPECTIVES FOR MANAGERS

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Abstract

Riparian vegetation is a central component of the hydrosystem. As such, it is often subject to management practices that aim to influence its ecological, hydraulic or hydrological functions. Remote sensing has the potential to improve knowledge and management of riparian vegetation by providing cost-effective and spatially continuous data over wide extents. The objectives of this review were twofold: to provide an overview of the use of remote sensing in riparian vegetation studies and to discuss the transferability of remote sensing tools from scientists to managers. We systematically reviewed the scientific literature (428 articles) to identify the objectives and remote sensing data used to characterize riparian vegetation. Overall, results highlight a strong relationship between the tools used, the features of riparian vegetation and the mapping extent. Very high-resolution data are rarely used for rivers longer than 100 km, especially when mapping species composition. Multi-temporality is central in remote sensing riparian studies, but authors use only aerial photographs and relatively coarse resolution satellite images for diachronic analyses. Some remote sensing approaches have reached an operational level and are now used for management purposes. Overall, new opportunities will arise with the increased availability of very high-resolution data in understudied or data-scarce regions, for large extents and as time series. To transfer remote sensing approaches to riparian managers, we suggest mutualizing achievements by producing open-access and robust tools. These tools will then have to be adapted to each specific project, in collaboration with riparian managers.

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1. INTRODUCTION

At the interface between terrestrial and aquatic biota, riparian vegetation is a central element in the hydrosystem, where it plays many ecological roles and interacts with all hydrosystem components (Naiman et al., 2005). In a broad sense, riparian vegetation corresponds to all vegetation types that grow within the area influenced by a river network (Naiman and Décamps, 1997).

Despite covering a relatively small area, riparian vegetation provides many ecosystem services related to river flow, sedimentary processes, river morphology, water chemistry, biodiversity and landscape identity (Dufour et al., 2019). However, riparian ecosystems experience multiple pressures (e.g. land use, water diversion, modified flood regime) and are threatened in many regions of the world (Stella and Bendix, 2019). Consequently, riparian vegetation is often the focus of management practices, including restoration or rehabilitation measures (Dufour and Piégay, 2009; González et al., 2015; Capon and Pettit, 2018), buffer implementation (Lee et al., 2004) or repeated maintenance operations such as wood removal (Piégay and Landon, 1997; Wohl et al., 2016).

In this context, management practices must be based on accurate and up-to-date information about the state of riparian vegetation. Many countries have thus established regional or national programs to monitor the health of riparian ecosystems (Bunn et al., 2010; Belletti et al., 2015). It is also necessary to sample riparian ecosystems densely when the objective is to plan and coordinate management practices (Landon et al., 1998; Beechie et al., 2008; Michez et al., 2017a) or assess their effectiveness (González et al., 2015). However, due to the spatial arrangement, dynamism and inaccessibility of riparian ecosystems, data acquisition in the field can be labor-intensive, especially for large areas (i.e. more than 100 km of a river) (Johansen et al., 2007). It is thus difficult to sample densely in the field, and the density or the extent of observations must be reduced. This can be problematic, because river scientists argue that small scale or discontinuous observations are inadequate to understand spatially continuous processes that occur at large spatial scales (Fausch et al., 2002; Marcus and Fonstad, 2008; see also Tabacchi et al., 1998 or Palmquist et al., 2018 for examples related to riparian vegetation).

Remote sensing techniques provide the ability to acquire continuous data over large extents. In the past few decades, the continued development of sensors, vectors and computational power has fueled the development of applications in environmental science (Anderson and Gaston, 2013; Wulder et al., 2012). In a recent literature review, remote sensing emerged as a particularly dynamic subject in riparian studies (Dufour et al., 2019). The use of remote sensing to study riparian vegetation generates specific challenges, especially related to the vegetation's relative structural complexity and spatial organization (Naiman and Décamps, 1997). Consequently, a wide array of data and methods have been used to address diverse contexts and objectives. For example, studies have focused on extracting specific features or processes related to riparian vegetation functions (e.g. surface roughness by Straatsma and Baptist (2008), shading of streams by Loicq et al., 2018). However, it is difficult to know whether and which remote sensing methods are relevant to a particular situation (Dufour et al., 2012). Many reviews and summary articles have addressed the use of remote sensing for mapping riparian vegetation, whether specifically or as part of a broader subject (Muller et al., 1993; Goetz, 2006; Ashraf et al., 2010; Dufour et al., 2012; Dufour et al., 2013; Forzieri et al., 2012; Dufour et al., 2019; Tomsett and Leyland, 2019). However, a recent, systematic and comprehensive assessment of the relevance of remote sensing approaches for mapping riparian vegetation is lacking.

Moreover, information is lacking about the transferability of remote sensing techniques to river managers who supervise operations to protect or restore riparian vegetation. This issue appears in many articles that address environmental management (Vanden Borre et al., 2011; Kennedy et al., 2009) or, more specifically, river management (Carbonneau and Piégay, 2012). This is not only a theoretical issue, since managers also raise this

issue. For example, this issue was discussed in a workshop that brought together managers and scientists to examine the remote sensing of fluvial corridors (Vivier et al., 2018). Moreover, many publications on the subject target non-academic stakeholders. For example, the technical report “New technologies in wetland studies: examples of applications” (Fédération des Conservatoires d’espaces naturels, 2018) reviews several examples of applications in academic and non-academic institutions, mainly in France. These articles and reports show that both scientists and managers are interested in using remote sensing to support riparian management.

The aims of this article are 1) to provide a comprehensive overview of the relevance of remote sensing to support the study and management of riparian vegetation and 2) to discuss the transferability of existing remote sensing techniques for managing riparian vegetation. To these ends, we first systematically review the different types of data used to study major features, functions and processes related to riparian vegetation across scales (sections 2 to 4). The second part of the review (section 5) qualitatively discusses the potential of using remote sensing tools to manage riparian vegetation and how these tools can be transferred from scientists to managers.

2. MATERIALS AND METHODS

2.1. DATABASE COLLECTION

Relevant articles were selected from the Scopus database (<https://www.scopus.com/>) for the period 1980 - April 2018, when the database was queried. We searched the title, abstract and the keywords for words related both to riparian vegetation and to remote sensing technologies. More precisely, we used the following request:

$$\left(\left(\begin{array}{l} \textit{riparian} \\ \textit{alluvial} \\ \textit{floodplain} \\ \textit{riverine} \end{array} \right) \textit{ within 5 words of } \left(\begin{array}{l} \textit{vegetation} \\ \textit{forest} \\ \textit{wood} \\ \textit{species} \\ \textit{tree} \end{array} \right) \right) \wedge \left(\left(\begin{array}{l} \textit{remote sensing} \\ \textit{multispectral} \\ \textit{hyperspectral} \\ \textit{hyperspatial} \\ \textit{SAR} \\ \textit{radar} \\ \textit{LiDAR} \\ \textit{UAV} \\ \textit{UAS} \\ \textit{drone} \\ \textit{photogrammetry} \end{array} \right) \wedge \left(\left(\begin{array}{l} \textit{imag}^* \\ \textit{photo}^* \\ \textit{data} \\ \textit{sensor} \end{array} \right) \textit{ within 5 words of } \left(\begin{array}{l} \textit{satellite} \\ \textit{thermal} \\ \textit{aerial} \\ \textit{spaceborne} \\ \textit{airborne} \end{array} \right) \right) \right)$$

Our choice of keywords excluded articles that mention riparian zones, but not specifically riparian vegetation. While some of these articles could have been relevant for this review, including keywords related to riparian zones would have resulted in unmanageable noise.

This request yielded 791 articles. We first filtered out irrelevant articles based on their title (672 articles kept). Then, we sorted through the remaining articles based on their abstracts (428 articles kept). During these two filtering steps, we removed mainly articles in which riparian vegetation was not an essential part of the study. For example, we removed geomorphological articles in which riparian vegetation was mentioned in the abstract but was not actually studied. Articles that used GIS but no remote sensing data were also removed (e.g. those using cadastral archives).

2.2. ANALYSIS GRID

We searched for features that characterized the articles collected to perform quantitative analysis and statistics. We built our analysis grid (Table 1) around five groups of variables: “general information”, “remote sensing technology”, “study extent”, “type of indicator” and “multi-temporality”. In the following section, when not obvious, we highlight in bold the codes (used in figures) associated with the variables. “General information” included variables such as the articles’ authors, title and location of study area. “Remote sensing technology” described the type of remote sensing data used. To simplify interpretation, we recorded this information as common combinations of sensors and vectors. We distinguished the following: airplane with a RGB/GS (red-green-blue or panchromatic), digital or analog sensor (**Plane_RGB**); airplane with a multispectral or hyperspectral sensor (**Plane_MSHS**); UAV with any sensor (**UAV**); any vector with a LiDAR sensor (**LiDAR**); any vector with a RADAR sensor (**RADAR**) and satellite with a multispectral or hyperspectral sensor. This last variable was coded according to image resolution: medium (> 10 m, **satlow**) or high (≤ 10 m, **sathi**). “Study extent” described the extent of the study area as the **length** of studied river or **area** of the study area. These two variables were recorded in categories and then summarized into a single category to simplify interpretation: **study extent**. “Type of indicator” described the type of features extracted with remote sensing data to describe riparian vegetation. Delineation of riparian vegetation among other land cover types (**DLC**) is the first feature extracted for managing riparian vegetation. Species composition is a major feature of riparian plant formations. It is related to habitat provision, bank stabilization and flood regulation functions; for example, willow is a pioneer species that helps to stabilize banks (Hupp, 1992). We distinguished studies that differentiate groups of species (**Communities**) and species (**SP**). We also distinguished studies in which the target species were invasive (**SP_invasive**), since riparian zones are particularly prone to invasions (Richardson et al., 2007). We distinguished studies in which the target communities were **succession stages**, since riparian systems are pulsed systems in which succession is regularly reinitiated, leading to a mosaic of succession stages (Kalliola and Puhakka, 1988). The structure of riparian vegetation is related to many ecological functions. We recorded general descriptors of vegetation structure such as vegetation **height**, **density**, **biomass** and **landscape** structure. We also recorded studies interested in hydraulic properties of vegetation (**Roughness**), since riparian vegetation has tremendous effects on the hydraulic regime of rivers, especially by slowing river flow (Curran and Hession, 2013). Riparian **shade** (or overhang) influences fish habitats and is a major factor regulating stream temperature (Poole and Berman, 2001). Large woody debris (**LWD**) has many effects on provision of aquatic habitats, river morphology and flood risk prevention (Wohl, 2017). Features related to physiological processes, including **phenology** and **health statuts** (e.g. tree dieback), are a major concern for managers (Cunningham et al., 2018). Riparian **evapotranspiration** has often been studied in arid or semi-arid systems because it has a major effect on providing water for human use (Dahm et al., 2002). “Multi-temporality” included only one variable (**dyna**), which corresponded to a special type of study – diachronic analysis – that uses a temporal series of images to describe vegetation dynamics. We recorded all variables as presence/absence data to capture the use of several types of data or the mapping of several indicators in the same article.

TABLE 1. ANALYSIS GRID USED FOR EACH ARTICLE IN THE DATABASE

Group of variables		Variable	Values	Description
General information		Author		
		Year		Publication year
		Journal		
		Country		Country of the study area
		X1		Longitude of the study area
		Y1		Latitude of the study area
		Biome		World Wildlife Fund Biome of the study area (extracted from the geographical coordinates of the study area)
Type of remote sensing data		Plane_RGB	0/1	Use of black and white or true-color aerial images (except images acquired from UAVs)
		Plane_MSHS	0/1	Use of aerial images with 4 or more spectral bands (except those from UAVs)
		Satlow	0/1	Use of satellite images with resolution > 10 m
		Sathi	0/1	Use of satellite images with resolution ≤ 10 m
		UAV	0/1	Use of images acquired from UAVs
		LiDAR	0/1	Use of LiDAR data
		RADAR	0/1	Use of RADAR data
Extent of the study area		Length	1 to 5	Length of the river studied (for studies at the scale of the minor bed or floodplain)
		Area	1 to 5	Area of the study area (for studies at the watershed scale)
		Study extent	1 to 5	Combination of Length and Area: Local: Length < 10 km River segment: Length 10-100 km OR Area < 100 km ² Subregional: Length 100-1000 km OR Area 100-1000 km ² Regional: Length > 10,000 km OR Area 1000-10,000 km ² Very large scale: Area > 10,000 km ²
Type of indicator	Delimitation	DLC	0/1	Mapping of riparian vegetation (including land cover studies)
	Species composition	Communities	0/1	Mapping of several distinct riparian plant communities
		Succession stages	0/1	Mapping of several succession stages
		SP	0/1	Mapping of riparian vegetation at the species level
		SP_invasives	0/1	Mapping of invasive species
	Vegetation structure	Height	0/1	Mapping of vegetation height
		Landscape	0/1	Calculation of landscape metrics (e.g. continuity)
		Density	0/1	Mapping of vegetation density
		Shade	0/1	Mapping of shade cast by vegetation
		Biomass	0/1	Mapping of biomass
		LWD	0/1	Large woody debris (wood in rivers)
		Roughness	0/1	Mapping of vegetation hydraulic properties
	Physiological processes	Evapotranspiration	0/1	Estimate of vegetation evapotranspiration
		Health status	0/1	Mapping of vegetation health status (e.g. tree dieback, defoliation)
Phenology		0/1	Mapping of vegetation phenology	
Multi-temporality		Dyna	0/1	Diachronic analysis

2.3. STATISTICAL ANALYSIS

We analysed the data collected with basic statistics and plots. Moreover, we performed a multiple correspondence analysis in order to highlight relationships between the type of data and the type of feature extracted. We used the package FactoMineR of R software. All variables were recorded as categorical variables. Variables related to study extent and multi-temporality were added as supplementary variables.

3. RESULTS

3.1. LOCATION OF THE STUDIES

Most studies in the 428 selected layed in the Northern Hemisphere, especially in North America (40% of studies) and Europe (20% of studies) (Figure 1). South America, Oceania, Asia (mostly Japan) and Africa represented respectively 9%, 9%, 11% and 5% of studies. Most represented biomes (Figure 1) were hardwood and mixed temperate forests (28%), temperate coniferous forests (14%), and deserts and xeric bushes (13%). Mediterranean biomes (10%) and temperate open biomes (8%) are also well represented. Well-represented biomes generally corresponded to those in developed countries. Conversely, boreal forests and tundra were least represented (< 1% of studies), though they cover a large area globally (> 10% of emerged land area). In addition, despite the large extent of tropical biomes (tropical and equatorial forests or open vegetation, ca. 30% of emerged land area), few studies focused on them.

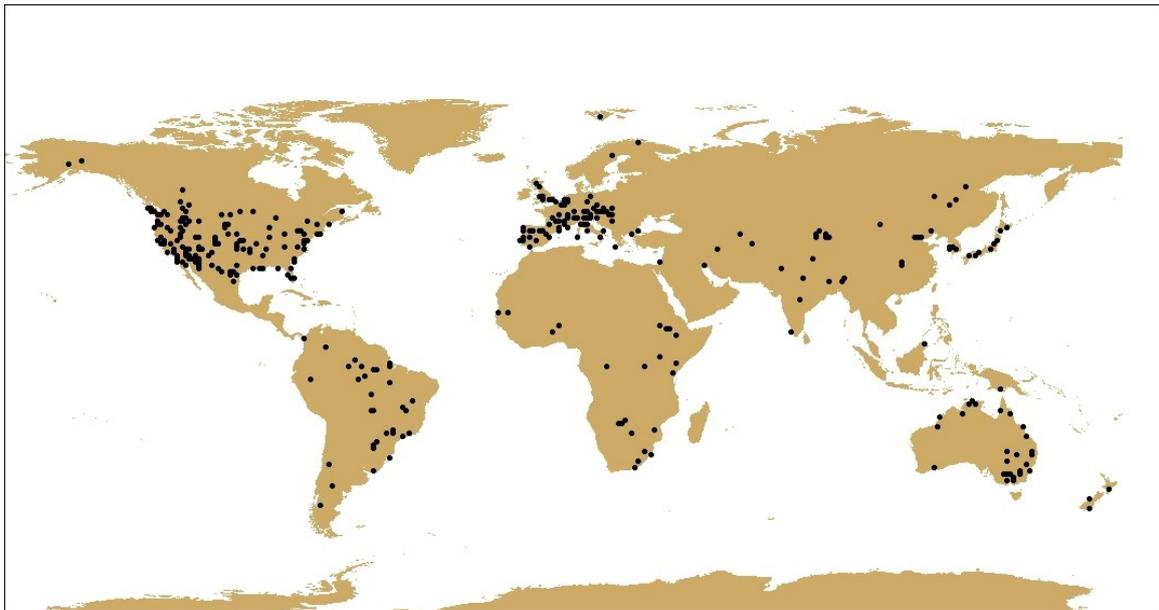


FIGURE 1. LOCATIONS OF THE STUDY AREAS OF THE STUDIES REVIEWED

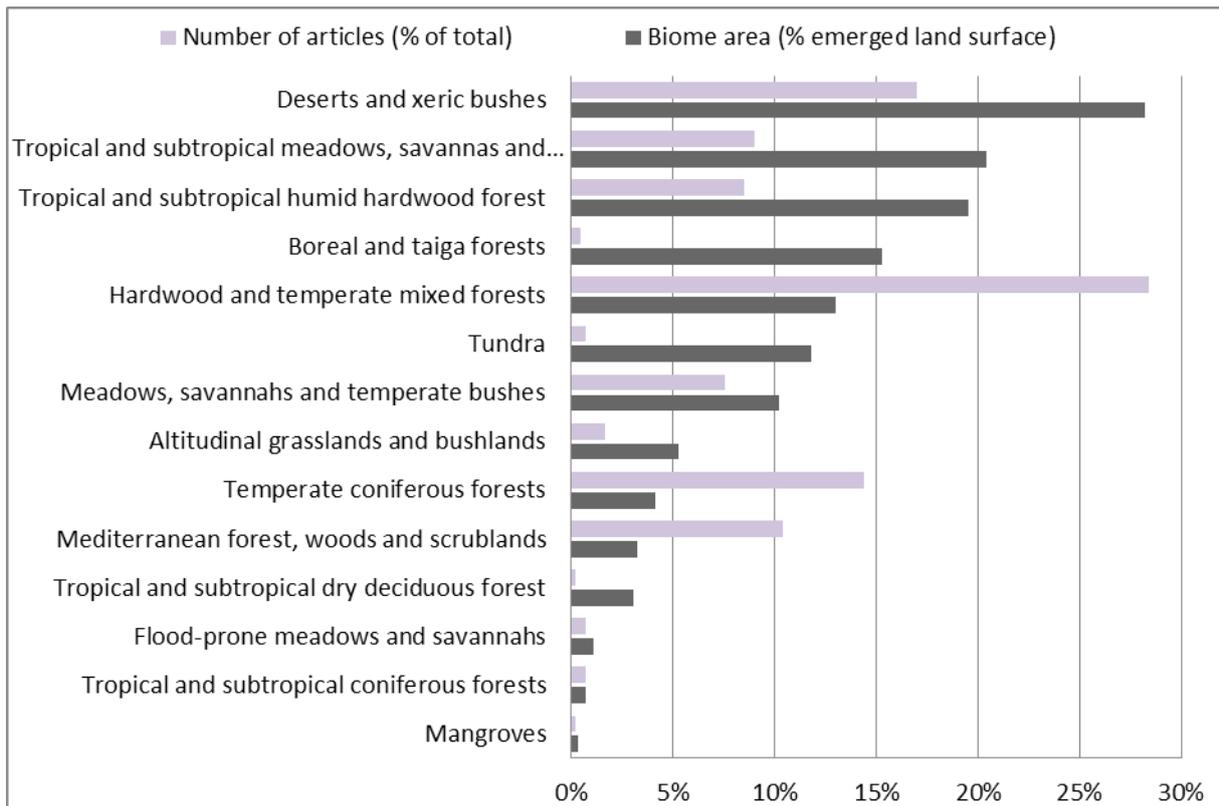


FIGURE 2. LOCATIONS OF STUDIES REVIEWED, BY WORLD WILDLIFE FUND BIOME

This result highlights the lack of knowledge and studies about tropical and boreal riparian forests, perhaps due to the location of laboratories, which are often located in developed countries and temperate climates. Our results are similar to those of Dufour et al. (2019) for all riparian vegetation studies and those of Bendix and Stella (2013) for studies of vegetation/hydromorphology relationships. The latter indicated that the most studied biomes were temperate forest, followed by semi-arid and desert shrubland, and Mediterranean forest. They also found a relationship between the main topic and the study area (e.g. the water table was studied more often in arid and semi-arid regions).

3.2. CHANGES OVER TIME IN THE NUMBER OF STUDIES THAT USED REMOTE SENSING TO STUDY RIPARIAN VEGETATION

Most of the 428 studies that used remote sensing to study riparian vegetation from 1980-2018 were published after 2000 (- left), when the number of studies began to increase greatly. Before 1990, few studies used remote sensing to study riparian vegetation. The percentage of studies using remote sensing among studies studying riparian vegetation increased in the 2000s (- right). After 2000, 2-6% of all studies of riparian vegetation used remote sensing. Thus, even recently, relatively few studies use remote sensing data to study riparian vegetation, and field-based approaches dominate riparian vegetation studies despite the development of remote sensing and modeling approaches. This could be due to three main reasons. First, field-based approaches have traditionally been used and are straightforward. Some aspects of riparian vegetation, such as biogeochemical functioning and soil properties, cannot realistically be studied with remote sensing (Dufour et al., 2012). Second, the spatial structure of riparian vegetation makes it difficult to study using remote sensing. Its complexity (Naiman et al., 2005) and narrow shape is difficult to observe with raw-resolution satellite images (Johansen et al., 2010). Additionally, the linear shape of riparian corridors requires acquiring images over large areas (to cover sufficient corridor length), only to focus on small areas (near the river, rather than other land-cover classes). For example, Weissteiner et al. (2016) estimated that Europe's riparian area represented ca. 1% of its total continental area. Third, we removed duplicate and irrelevant articles from our

database, but did not do so when identifying all articles describing studies of riparian vegetation in general, which may have led us to underestimate the percentage of all riparian studies that used remote sensing.

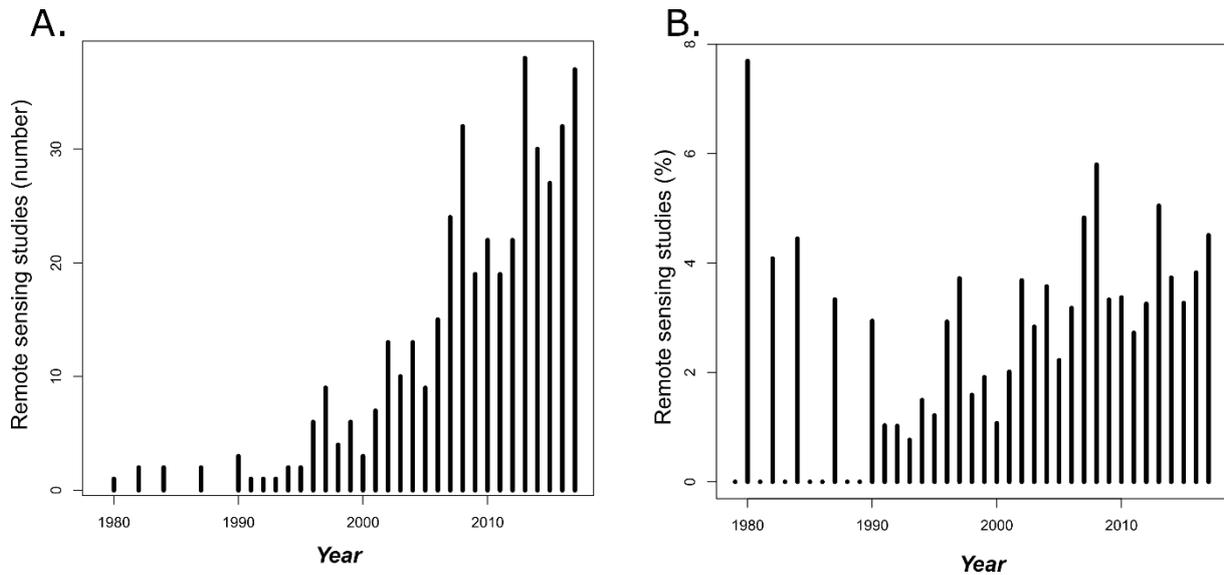


FIGURE 3. A. NUMBER OF STUDIES FROM 1980-2018 THAT USED REMOTE SENSING TO STUDY RIPARIAN VEGETATION. B. PERCENTAGE OF STUDIES FROM 1980-2018 THAT USED REMOTE SENSING, OUT OF ALL STUDIES OF RIPARIAN VEGETATION (IDENTIFIED USING THE SAME KEYWORDS AS THOSE IN THE MAIN DATABASE, WITHOUT THOSE RELATED TO REMOTE SENSING).

3.3. CHANGES IN REMOTE SENSING TECHNIQUES OVER TIME

The remote sensing data used most were aerial RGB/GS images (44% overall) and medium-resolution satellite images (> 10 m resolution, and ≤ 50 m for most studies) (Figure 3). Aerial multispectral images appeared in the 1990s and peaked during the 2000s. The use of high-resolution satellite data (≤ 10 m), such as IKONOS, SPOT 5 and WorldView, started in the late 1990s and reached a plateau around 2010. The use of LiDAR data consistently increased during the 2000s, accounting for 20% of studies using remote sensing for riparian vegetation in 2017. The use of UAV images sharply increased in the 2010s. As the use of these technologies increased, the percentage of studies using RGB/GS aerial images decreased slightly.

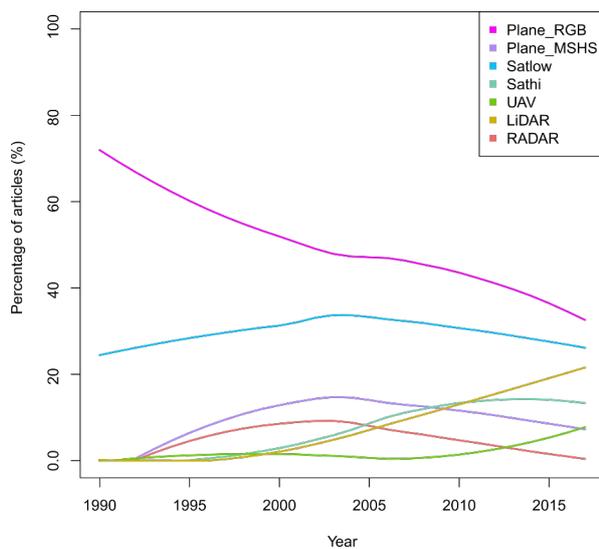


FIGURE 4. PERCENTAGE OF STUDIES FROM 1980-2000 THAT USED A GIVEN TECHNOLOGY PER YEAR. THE CURVE WAS SMOOTHED USING A LOESS REGRESSION.

Conversely, less than 2% of studies used RADAR data. Its use peaked in the early 2000s and then decreased, being replaced with LiDAR data. This low percentage of RADAR data and its replacement by LiDAR data could be due to the lower spatial resolution of RADAR sensors than LiDAR sensors (generally < 1m), the relative difficulty of interpreting RADAR data and because water surfaces can modify RADAR signals and make them even more difficult to interpret. Most studies in our database that used RADAR data focused on the interaction between water and riparian vegetation, mapping flooding events or roughness coefficients (Townsend, 2002). Consequently, this type of data seems less appropriate for studying riparian vegetation than optical images (the main satellite sensor) or LiDAR data, another type of sensor that retrieves structural information.

3.4. MULTI-TEMPORALITY OF REMOTE SENSING RIPARIAN STUDIES

Overall, 54% of studies in the database were multi-temporal (i.e. studies where in which data acquired at on several dates are used to understand the dynamics of riparian vegetation). RGB/GS aerial images were used in more than 60% of the multi-temporal studies (Figure 5a), such as those of Lallias-Tacon et al. (2017) and Dufour et al. (2015). This may have been because this type of image is simple to use and has been available over a large extent since the 1950s (Dufour et al., 2012). In most of the countries previously highlighted as active in riparian research, public administrations have performed long-term and systematic national aerial surveys for general purposes (e.g. urban planning) that researchers can use at low cost. Most multi-temporal studies that included aerial photographs used photointerpretation to describe riparian vegetation features.

Conversely, more recent technologies (e.g. high-resolution satellite images, LiDAR data) were far more common in studies that focused on one period than in multi-temporal studies (Figure 5b). For example, LiDAR and high-resolution satellite data were used in respectively 24% and 18% of “mono-temporal” studies vs. 4% and 5% of multi-temporal studies. Mono-temporal studies included methods developed to map riparian forest attributes with remote sensing data. In these studies, the methods used were more complex and mostly automated, such as supervised classifications (Michez et al., 2016b; Antonarakis et al., 2008) and calculation of metrics (Riedler et al., 2015).

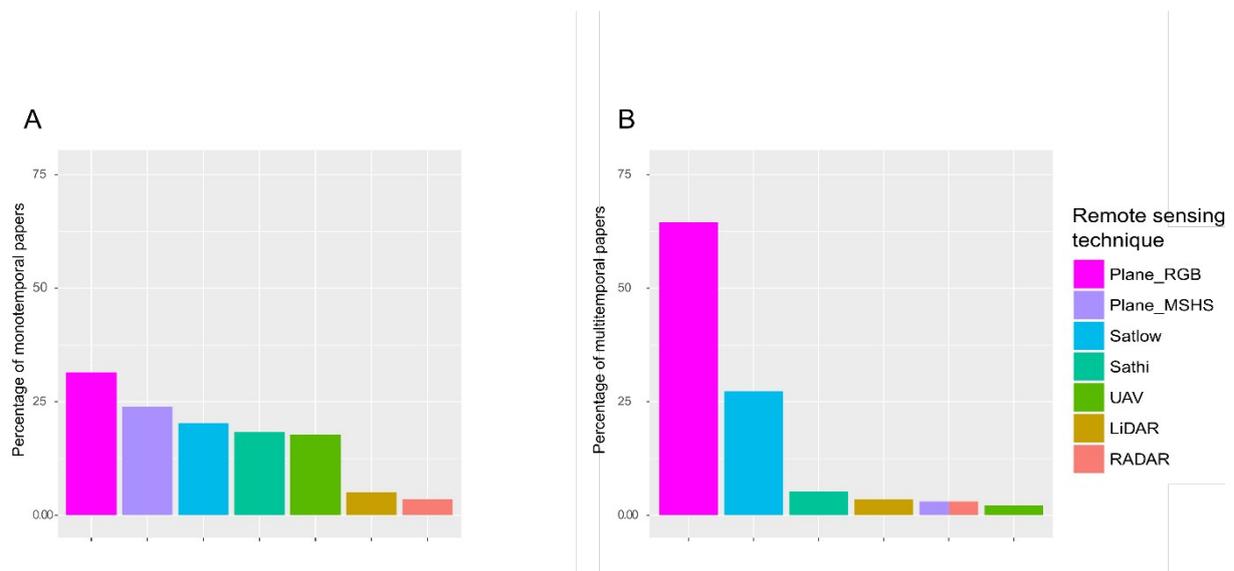


FIGURE 5. USE OF REMOTE SENSING DATA IN (A) MULTI-TEMPORAL AND (B) MONO-TEMPORAL STUDIES (RESPECTIVELY 54% AND 46% OF THE STUDIES).

3.5. WHICH TECHNOLOGY FOR WHICH STUDY SCALE?

There was a strong relationship between the scale of the study (landform to river basin) and the type of remote sensing data used (Figure 6). In general, aerial images were used more at relatively local scales (i.e. local and river segment), while medium-resolution satellite images were used more at larger scales (i.e. regional or very large scale). There was often a tradeoff between resolution and coverage: UAVs can produce images with centimetric resolution but struggle to cover large areas, while satellites such as Landsat and MODIS provide images at a lower resolution (30 m for Landsat, 250 m for MODIS) but can cover large areas.

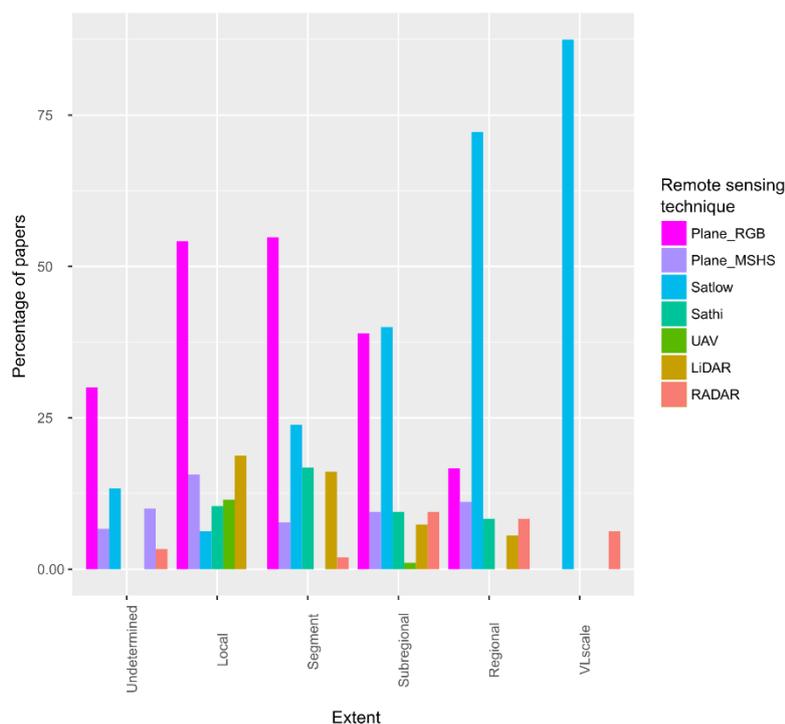


FIGURE 6. PERCENTAGE OF STUDIES THAT USED A GIVEN REMOTE SENSING TECHNOLOGY, BY SPATIAL EXTENT OF THE STUDY

3.5.1. LOCAL SCALE

At the local scale (< 10 km long), 86% of studies were based on airborne remote sensing (of which 79% used airplanes and 11% used UAVs). This scale of study lies within the range of action of relatively inexpensive UAVs that can carry RGB and multispectral cameras. While most UAVs were used at the local scale, the relatively low percentage local-scale studies that used UAVs was surprising. This could have been due to the recent availability of these platforms: of studies published in the 2010s, 20% of those at the local scale used UAVs. UAVs are considered more versatile than planes, and a growing number of “ready-to-fly” platforms allow end-users to perform their own acquisitions (Anderson and Gaston, 2013). Moreover, UAV imagery provides very high spatial resolution imagery (up to centimetric), which is ideal for operator photointerpretation, which is frequently used at this scale. However, most developed countries have established regulations that restrict the potential and spread of UAV technology (Stöcker et al., 2017).

3.5.2. INTERMEDIATE SCALES: THE DIFFICULTY IN UPSCALING HIGH-RESOLUTION DATA

Both airborne and spaceborne sensors were used at intermediate scales (10-1000 km or < 10,000 km²). RGB/GS aerial images were used in 55% and 39% of studies at respectively the river-segment (10-100 km) and subregional scale (100-1000 km). Most researchers photointerpret these images to describe riparian vegetation

features. This method is long-standing, but remains a relevant and effective approach to map riparian vegetation over small watersheds or along dozens (more rarely hundreds) of km of rivers (Jansen and Backx, 1998; Matsuura and Suzuki, 2013; Carli and Bayley, 2015; González del Tánago et al., 2015; Solins et al., 2018). However, photointerpretation of hundreds of km of river can become tedious. In this case, one would use more automated approaches, such as object-based approaches, which can decrease the time required for photointerpretation (Belletti et al., 2015a).

However, the effectiveness of automated techniques is strongly correlated with the homogeneity of spectral signatures within a single feature class. Homogeneity in spectral signatures requires homogeneous atmospheric and illumination conditions within the dataset. To this end, airplanes can be used over long river segments in a short period to avoid variations in weather and illumination conditions (Forzieri et al., 2013; Bucha and Slávik, 2013). However, this approach remains challenging for large river networks, which decreases the possibility of automation at these scales (Dauwalter et al., 2015). In this context, the wider swath of satellite imagery would be an advantage. High-resolution satellite images were often used to map vegetation automatically (16% and 9% of studies at respectively the river-segment and subregional scale). For example, Strasser and Lang (2015), Riedler et al. (2015) and Doody et al. (2014) used WorldView-2 data to map riparian vegetation along a few dozen km. Tormos et al. (2011) and Macfarlane et al. (2017) used SPOT images and GeoEye-1 images to map vegetation along corridors respectively 60 and 90 km long. However, it may be difficult to acquire high-quality datasets for larger areas, for which several high-resolution satellite images must be combined (Goetz, 2002; Johansen et al., 2010b; Zogaris et al., 2015).

The percentage of studies based on LiDAR surveys decreased with scale: 19%, 16%, 7% and 6% of studies at respectively the local, river-segment, subregional and regional scale. However, some authors were able to use LiDAR data to monitor narrow riparian corridors over large areas (Johansen et al., 2010b; Michez et al., 2017b). One advantage of tri-dimensional LiDAR data is that they are less subject to changing atmospheric and lightning conditions during the survey than spectral data. Moreover, LiDAR coverage is becoming more frequent at the regional/national scale (Tompalski et al., 2017; Parent et al., 2015; Wasser et al., 2015; Shendryk et al., 2016). When an initial nationwide LiDAR survey is performed, digital aerial photogrammetry (DAP) can be used to further update LiDAR canopy-height models (CHMs). DAP CHMs can be produced from aerial images acquired on a regular basis by national or regional mapping agencies in several countries and can potentially provide quality vegetation height information at low additional cost (Michez et al., 2017b).

3.5.3. LARGE SCALE: SATELLITE IMAGES

The use of satellite images with medium to coarse resolution (> 10 m) increased as the extent increased. For studies at the regional or large scale, satellite images were used in respectively 72% and 82% of cases. Coarse-resolution images (> 100 m) were not used to study riparian vegetation, which often appears as linear or fragmented features (Gergel et al., 2007). Medium-resolution images such as Landsat TM and ETM+ images are preferred. However, the resolution of these data are still insufficient to study narrow riparian buffers (Goetz, 2006). Although aerial images (multispectral, RGB and panchromatic) were used in 25% of studies at the regional scale, they were always used with medium-resolution satellite images (Fullerton et al., 2006; Claggett et al., 2010; Groeneveld and Watson, 2008). High-resolution satellite images, which were used in 8% of studies at the regional scale, were used mostly with pansharpening methods to enhance lower-resolution satellite images (Staben and Evans, 2008; Seddon et al., 2007; Scott et al., 2009).

3.6. WHICH TECHNOLOGY FOR WHICH RIPARIAN FEATURE?

The features of interest extracted from remote sensing data to describe riparian vegetation were strongly related to the type of remote sensing data (Figure 7). Four major trends/groups emerged. First, the study of physiological processes (e.g. phenology, evapotranspiration and, to a lesser extent, health status) was strongly associated with the use of medium-resolution satellite images and large study extents (Fig. 6, left). Second, the study of features or processes related to vegetation structure (shade, roughness, height) was strongly

associated with the use of LiDAR data (Fig. 6, top right). The study of features related to species composition was associated with the use of high-resolution multispectral images (acquired from satellites, planes or UAVs) or RGB/GS aerial images (especially for successional stages) and with small study extents (Fig. 6, bottom right). The delineation of riparian vegetation was weakly associated with the use of RGB/GS aerial images or medium-resolution satellite images (Fig. 6, center). These four trends are discussed in the following four sections

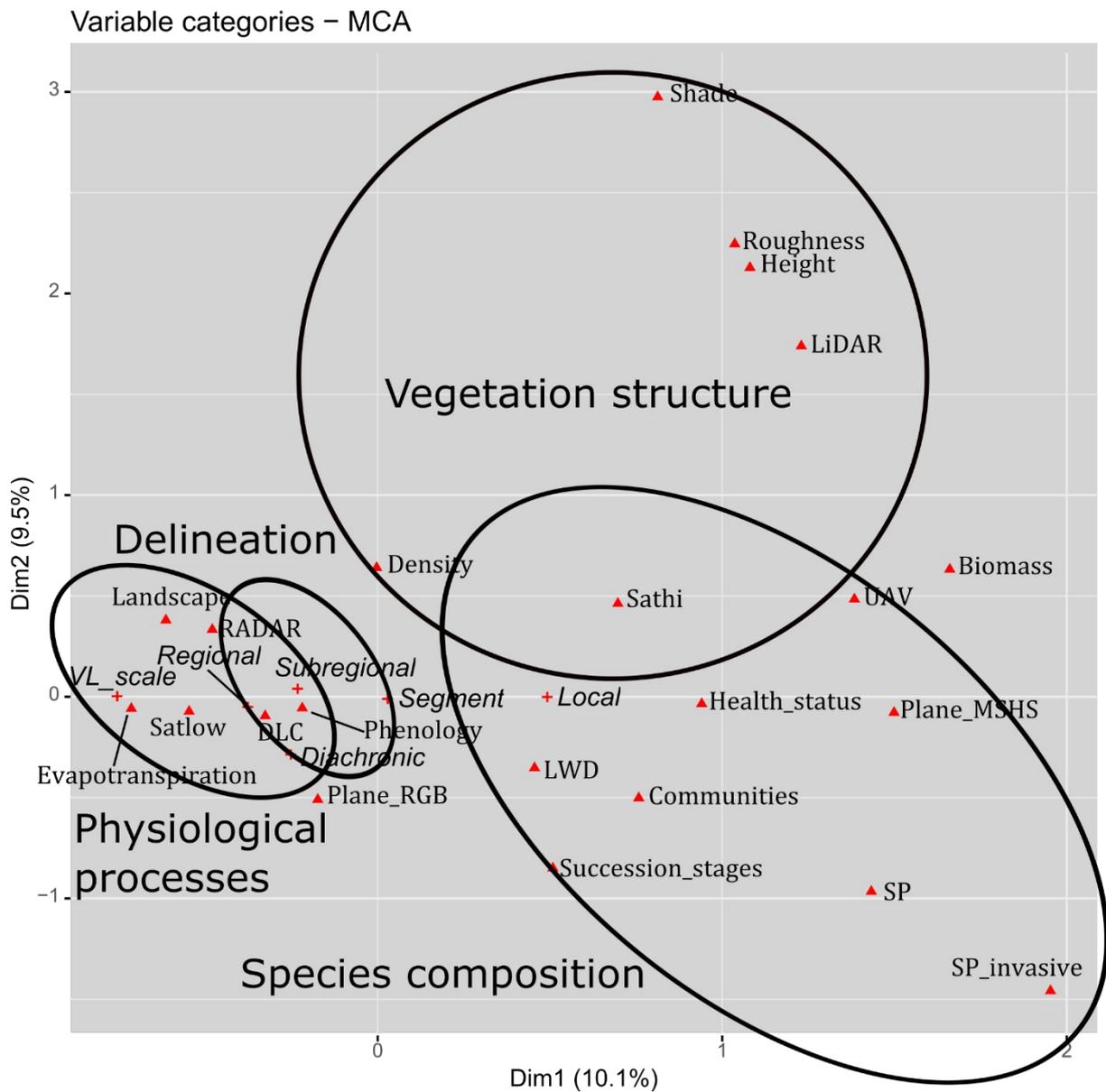


FIGURE 7. RESULTS OF THE MULTIPLE CORRESPONDENCE ANALYSIS (SEE SECTION 2.3. FOR THE METHODS). SUPPLEMENTARY VARIABLES (I.E. VARIABLES RELATED TO STUDY EXTENT AND MULTI-TEMPORALITY) ARE REPRESENTED AS CROSSES WITH TEXT IN ITALICS. THE FIRST TWO AXES EXPLAIN 19.6% OF TOTAL VARIANCE. ELLIPSES WERE DRAWN ARBITRARILY TO SIMPLIFY INTERPRETATION. SEE TABLE 1 FOR CODE DEFINITIONS.

3.6.1. DELINEATION OF RIPARIAN VEGETATION

How riparian vegetation is delineated depends on how it is defined (Verry et al., 2004). In general, riparian vegetation is defined based on its specific characteristics (e.g. spectral signature, texture) and on contextual information (e.g. topographic position, proximity to a river) (Weissteiner et al., 2016). Photointerpretation of RGB/GS aerial images is a traditional approach in which the operator uses both types of information (Morgan et al., 2010). It was used in 53% of studies that delineated riparian vegetation. Multispectral images (airborne or spaceborne, accounting for 45% of studies) are often used to delineate riparian vegetation in an automated way (Alaibakhsh et al., 2017; Johansen et al., 2010b; Bertoldi et al., 2011) (Figure 7B). Contextual information can be provided by ancillary data (e.g. hydrographic network, as in Claggett et al. (2010) or Yang (2007)), a LiDAR digital terrain model (DTM) (Arroyo et al. 2010; Wagner-Lücker et al. 2013), or a Shuttle RADAR Topography Mission DTM (Maillard and Alencar-Silva, 2013; Weissteiner et al., 2016). Congalton et al. (2002) indicate that medium-resolution satellite data (used in 29% of studies) are not adapted for delineating narrow riparian corridors because the corridors do not contain enough pixels.

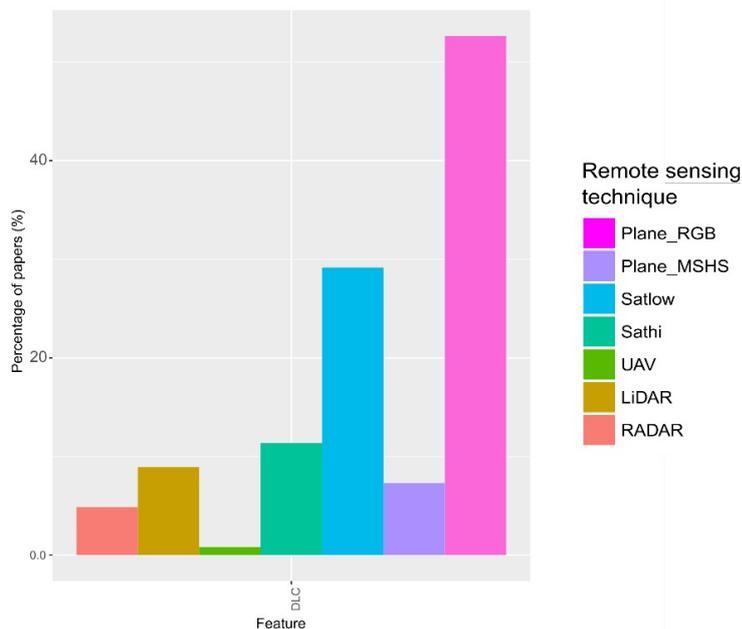


FIGURE 7B. PERCENTAGE OF STUDIES THAT USED GIVEN REMOTE SENSING DATA TO DELINEATE RIPARIAN VEGETATION (I.E. DISTINGUISH RIPARIAN VEGETATION FROM OTHER LAND-COVER TYPES)

3.6.2. SPECIES COMPOSITION

Species composition is a recurrent subject that was studied in 42% of studies (including 8% and 7% of those interested respectively in successional stages and invasive species). Photointerpretation of RGB/GS aerial images concerned 51%, 47% and 45% of studies that differentiated respectively communities, species, and invasive species. This approach is widely used to describe successional stages or changes in their distribution (86% of such studies). Indeed, RGB/GS aerial images have been available since before the 1950s (González et al., 2010; Rood et al., 2010; Varga et al., 2013; Wan et al., 2015). However, manual interpretation of images is time-consuming, and the discriminating power of RGB/GS aerial images is limited by their low spectral range (Narumalani et al., 2009; Fernandes et al., 2014). Medium-resolution satellite images were used in 21% of studies that differentiated communities. These images were used mainly when vegetation patches were larger than the image resolution (Vande Kamp et al., 2013; Hamandawana and Chanda, 2013; Sridhar et al., 2010; Groeneveld and Watson, 2008; Townsend and Walsh, 2001), although spectral unmixing can, to some extent, resolve this issue (Gong et al., 2015; Wang et al., 2013).

The most promising approaches to address this issue are based on high-resolution, aerial or spaceborne, multispectral or hyperspectral images, such as mono- or multi-temporal images. These images were used in 30%, 33% and 45% of studies that differentiated respectively communities, species and invasive species. The accuracy of a particular project depends on the context, objectives, available data and methods used to evaluate it. Therefore, we present recent studies that mapped species in the Table 2. In general, a large number of narrow spectral bands increases the ability to distinguish species. However, in mature, species-rich floodplain forests, it remains challenging to obtain classification accuracy that is satisfactory for operational use, even when using hyperspectral imagery (Richter et al., 2016). The use of multi-temporal images, which reveal the succession of phenological stages, can sometimes replace the spectral range (Michez et al., 2016b). It is also possible to acquire images at a single but appropriate date to take advantage of the singular aspect of one species at a particular phenological stage. This approach is especially effective when a single species has to be mapped, such as the invasive species *Arundo donax* (Fernandes et al., 2013b) or *Heracleum mantegazzianum* (Michez et al., 2016a). The spatial resolution of images must be sufficiently high to limit the occurrence of mixed pixels that hinder the performance of automated classifications (Belluco et al., 2006; Narumalani et al., 2009). However, small mixtures of species remains a source of difficulty, even with a cm resolution (Michez et al., 2016a). LiDAR data, also used to classify species, can supplement multispectral data to provide information about vegetation height (Forzieri et al., 2013) or to segment trees before classifying them (Dutta et al., 2017). They have also been used as the sole source of data by relating species identity to the structure of the point cloud (Laslier et al., 2019).

TABLE 2. EXAMPLES OF REMOTE SENSING METHODS USED TO CLASSIFY RIPARIAN SPECIES IN DIFFERENT SETTINGS AND THEIR ACCURACY

Reference	Data	Classes	Accuracy	Comment
Mature riparian forests				
Fernandes et al. (2013a)	RGB-NIR aerial imagery (0.5 m resolution)	3 types of mature, temperate/Mediterranean riparian forests	61 (small) - 78% (large river)	
Dunford et al. (2009)	RGB imagery acquired with UAV (0.13 m resolution)	4 tree species (<i>Populus</i> , <i>Salix</i> and 2 <i>Pinus</i>) in a riparian Mediterranean forest	91% (for an image) - 71% (for a mosaic)	
Michez et al. (2016b)	RGB-NIR imagery acquired with UAV (0.1 m resolution)	5 tree species in a temperate, riparian forested/agricultural landscape	84 (forested) - 80% (agricultural)	Multi-temporal dataset
Richter et al. (2016)	Hyperspectral aerial imagery (367 bands, 2 m resolution)	10 tree species in a mature temperate floodplain forest	74% (single-date survey) - 78% (two-date survey)	
Dutta et al. (2017)	Hyperspectral aerial imagery (48 bands, 1 m resolution)	4 groups of tree species in a mature, temperate riparian forest	86%	LiDAR is used to segment the trees
Laslier et al. (2019)	High density (> 45 points/m ²) LiDAR point cloud	8 tree species in a temperate riparian agricultural/forested landscape	67%	
Pioneer/species-poor riparian settings				
Macfarlane et al. (2017)	Pansharpened GeoEye-1 imagery (RGB-NIR, 0.5 m resolution)	Pioneer (<i>Salix</i> , <i>Populus</i>) and invasive (<i>Tamarix</i>) species in an arid context	80%	
Forzieri et al. (2013)	RGB-NIR aerial imagery (0.2 m resolution); hyperspectral aerial imagery (102 bands, 3 m resolution) and LiDAR data (DSM/DTM with 1 m resolution)	Pioneer (<i>Salix</i> , <i>Populus</i>) and invasive (<i>Arundo donax</i>) species in a temperate context	93%	
Invasive species				
Narumalani et al. (2009)	Hyperspectral aerial imagery (62 bands, 1.5 m resolution)	<i>Tamarix</i> , <i>Elaeagnus angustifolia</i> , <i>Cirsium arvense</i> , <i>Carduus nutans</i> and mixed classes	74%	Mixed classes are not well classified and decrease overall accuracy
Fernandes et al. (2014)	RGB-NIR aerial imagery (0.5 m resolution)	<i>Arundo donax</i>	97%	Choice of the best date for aerial survey
	WorldView 2 imagery (8 bands, 2 m resolution)	<i>Arundo donax</i>	95%	
Michez et al. (2016a)	RGB-NIR imagery acquired with UAV (0.05-0.1 m resolution)	<i>Impatiens glandulifera</i>	72%	Mixture with native species hinders accurate classification
		<i>Heracleum mantegazzianum</i>	97%	
		<i>Fallopia japonica</i>	68%	
Peerbhay et al (2016)	WorldView 2 imagery (8 bands, 2 m resolution)	<i>Solanum mauritanum</i>	68%	
Miao et al. (2011)	Hyperspectral aerial imagery (227 bands, 1 m resolution)	<i>Prosopis glandulosa</i> and <i>Tamarix</i>	92%	
Doody et al. (2014)	WorldView 2 imagery (8 bands, 2 m resolution)	<i>Salix</i>	93%	

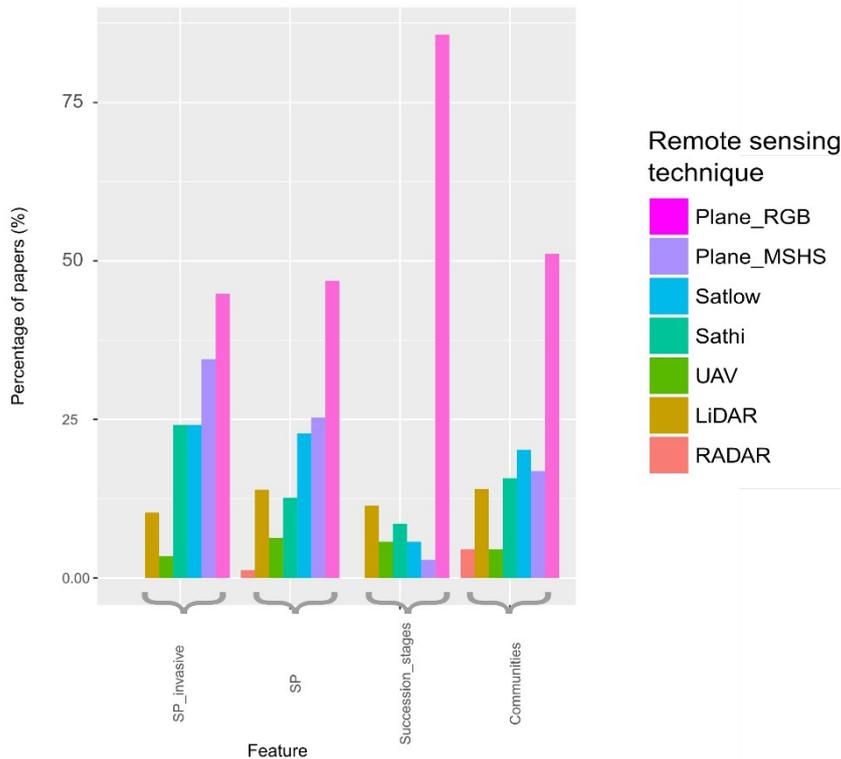


FIGURE 8. PERCENTAGE OF STUDIES THAT USED GIVEN REMOTE SENSING DATA TO MAP INDICATORS RELATED TO SPECIES COMPOSITION.

3.6.3. PHYSIOLOGICAL PROCESSES

Medium-resolution satellite images (> 10 m resolution and ≤ 50 m for most studies) were the most popular type of data used to assess physiological processes of riparian vegetation (100%, 73% and 54% of studies concerning respectively evapotranspiration, phenology and health status) (Figure 8B). One advantage of using these images in this context is that they are often available as dense series, which is useful for studying cyclic processes. For example, Wallace et al. (2013) used AVHRR images (return period < 1 day) to detect variations in the timing of greening up/scenescing of vegetation. Nagler et al. (2012) used MODIS (return period 1-2 days) to study phases of the life cycle of the tamarix leaf beetle (*Diorhabda carinulata*) throughout the year. Cadol and Wine (2017) and Nagler et al. (2016) used long-term records (several years) of satellite images along with flow data to investigate relationships between hydrology and physiological processes in riparian vegetation. Sims and Colloff (2012) used MODIS images over several years to assess responses of riparian vegetation during and after flooding events. However, the low resolution often means that pixels in the image aggregate greater heterogeneity in ground features; accuracy thus decreases, making it more complicated to study different types of vegetation separately (Tillack et al., 2014; Cunningham et al., 2018). The health status of vegetation is often studied with higher resolution data, and occasionally with a single image (Tillack et al., 2014; Michez et al., 2016b; Bucha and Slávik, 2013; Shendryk et al., 2016; Sankey et al., 2016).

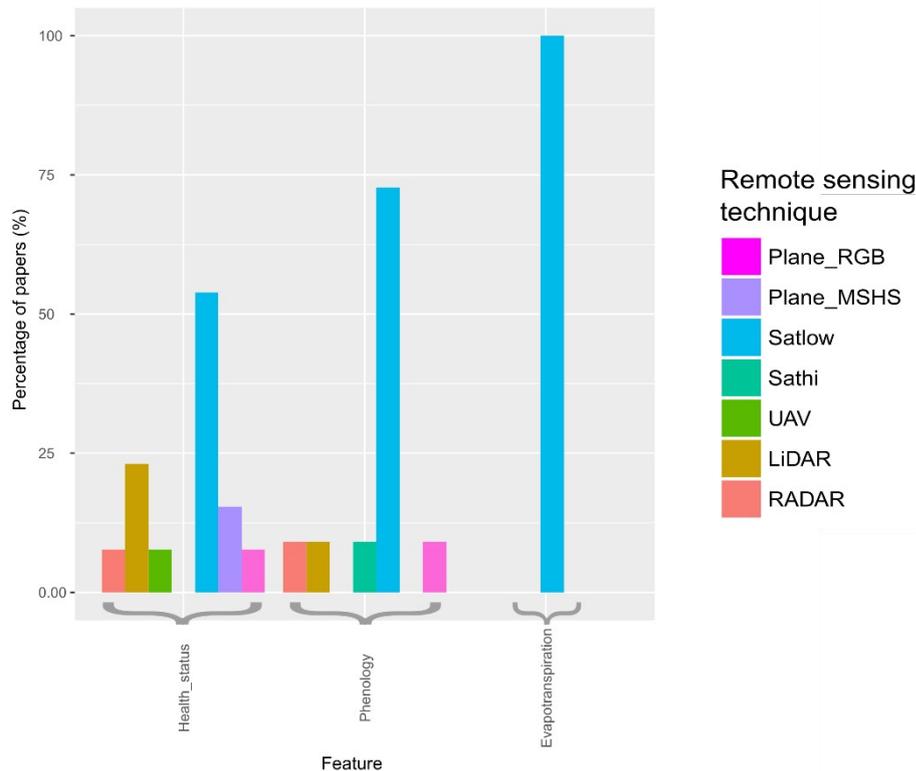


FIGURE 8B. PERCENTAGE OF STUDIES THAT USED GIVEN REMOTE SENSING DATA TO DESCRIBE PHYSIOLOGICAL INDICATORS.

3.6.4. VEGETATION STRUCTURE

LiDAR appears to be the most promising technology for describing vegetation structure and related functions (shading, surface roughness). The LiDAR signal can penetrate the canopy and the water surface, and provides information about topography under dense canopies, the internal structure of canopies and bathymetry. Retrieving simple structural attributes of vegetation (e.g. height, continuity, overhanging character) is straightforward, since they can be extracted from DTMs, DSMs or CHMs delivered by LiDAR data producers. Alternatively, DAP can be used to produce similar photogrammetric DSMs, which can be combined with LiDAR DTMs to produce updated high-quality CHMs at lower cost (Michez et al., 2017b).

These applications have reached an operational level. However, further methodological developments for processing the 3D point cloud and new generations of full-waveform LiDAR data must be explored before they can be transferred to management operations. For example, full-waveform LiDAR data have shown promising results in forestry applications (e.g. Koenig and Höfle, 2016), but there are few examples for riparian vegetation (Shendryk et al., 2016).

LiDAR data have often been used to map riparian shade, which is a major parameter that influences stream water temperature (Poole and Berman, 2001). Temperature regulates the habitat of aquatic species such as the brown trout (*Salmo trutta fario* L.) (Caissie, 2006; Georges et al., in press), and the effect of riparian shade on stream water temperature is strong enough to affect aquatic communities significantly (Bowler et al., 2012). Field methods used to measure stream shade are expensive and time-consuming (Rutherford et al., 2018). LiDAR data appears to be the most promising alternative because they can describe shade at a fine scale (Richardson et al., 2019). Several methods for using LiDAR data to measure riparian shade have been described in the literature. Richardson et al. (2009) calculated light penetration index raster products as a predictor of light conditions. LiDAR data can describe shadowing properties using a simple CHMs derived from point clouds (Michez et al., 2017; Loicq et al., 2018; Wawrzyniak et al., 2017). Other studies have used 3D point clouds to retrieve the finest-scale information about vegetation structure. For example, Akasaka et al. (2010) used a

LiDAR point cloud to estimate biomass overhanging the river, while Tompalski et al. (2017) used one to model solar shading on a given summer day. Recently, Shendryk et al. (2016) used full-waveform LiDAR data to estimate the dieback of individual riparian trees, which was related to their shadowing properties.

LiDAR data have also been used to map floodplain roughness in a spatially continuous manner. Forzieri et al. (2012) distinguished two main approaches for mapping floodplain roughness using remote sensing: classification-derived mapping and hydrodynamic modeling. In the former, thematic maps of land cover or vegetation classes are produced with remote sensing data. A roughness coefficient (often Manning’s coefficient) is then assigned to each class using a lookup table. In the latter, hydrodynamic properties of vegetation are estimated using an indicator of vegetation structure (e.g. leaf area index, stem or crown diameter, vegetation height). LiDAR technology has several advantages in this case: it measures structural attributes directly and can account for complex, multilayered structures (Manners et al., 2013; Jalonen et al., 2015). Hydrodynamic modeling is often combined with classification-derived mapping, with separate modeling of hydrodynamic properties of each vegetation class (Straatsma and Baptist, 2008; Zahidi et al., 2018). Development of restoration and multi-objective management practices (to promote ecosystem health while protecting people and goods) has increased demand for models that represent effects of vegetation on flow more accurately (Rubol et al., 2018). However, research on hydrodynamic properties of vegetation and how to measure them in the field is ongoing (Shields et al., 2017).

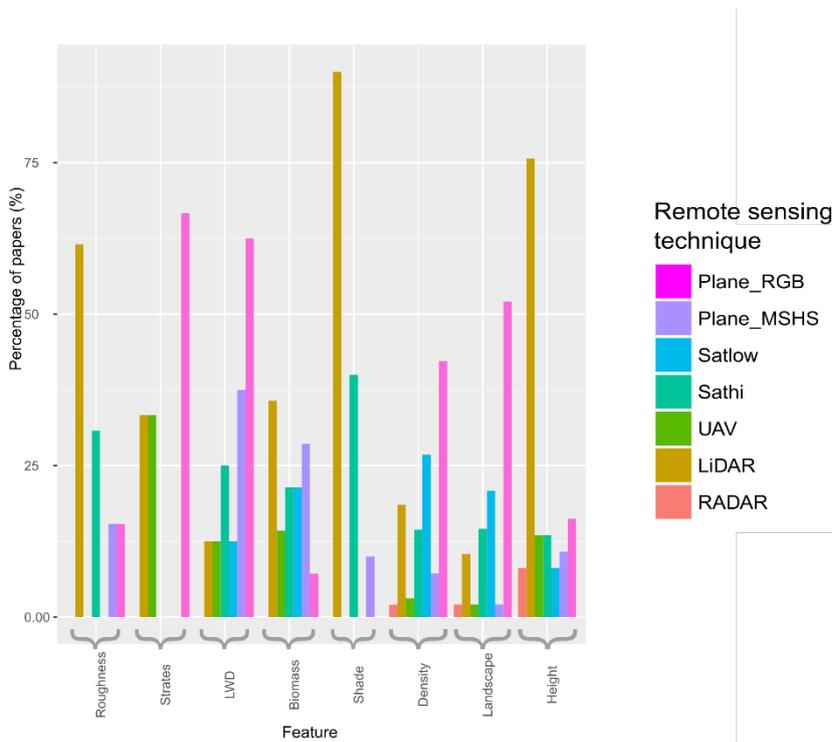


FIGURE 9. PERCENTAGE OF STUDIES THAT USED GIVEN REMOTE SENSING DATA TO MAP STRUCTURAL FEATURES OF RIPARIAN VEGETATION.

4. DISCUSSION: TRENDS AND GAPS EMERGING FROM THE DATABASE

The main perspectives that emerged from our results were the use of remote sensing to widen the geographical scope of riparian research, the new opportunities provided by denser time series acquired from new sensors and the challenge of scaling up riparian studies with high-resolution data.

4.1. ENLARGING THE GEOGRAPHICAL SCOPE OF RIPARIAN RESEARCH WITH REMOTE SENSING

We showed that certain regions of the world, despite their large extents, lack studies of riparian vegetation with remote sensing. This is true for boreal and tropical ecosystems, and for developing countries. Dufour et al. (2019) made the same observation for all types of riparian vegetation studies. However, we suggest that the increasing quality of remote sensing data has great potential for research in understudied areas and at the global scale. One condition is that these data must be available to their potential users. Open or free remotely sensed data, such as Landsat, MODIS or, more recently, Sentinel images, allow researchers to resolve the issue of the prohibitive cost of data acquisition. This is particularly true for researchers in developing countries for data that are produced in wealthier countries (Sá and Grieco, 2016). However, to broaden the user base, it is also necessary to facilitate access to these data (Turner et al., 2015). Access can be facilitated by providing higher-level (e.g. atmospherically corrected) or derived products, such as global land cover maps (Gong et al., 2013), global floodplain models (Nardi et al., 2019) and maps of riparian zones (Weissteiner et al., 2016, at the European scale). Access can also be made easier by developing an open, free or user-friendly environment to find, visualize and process data (Turner et al., 2015).

4.2. USING DENSER TIME SERIES ACQUIRED FROM NEW SENSORS

Our review showed that multi-temporality was a major feature of riparian vegetation studies (55% of studies were multi-temporal). Most multi-temporal studies were performed using aerial orthophotos and described inter-annual processes, mostly on decadal time scales. This may have been because until currently, this type of data was the easiest to acquire over a long time series. For example, aerial orthophotos have been available in France since the 1950s. Until recently, campaigns for acquiring aerial images were performed relatively infrequently, resulting in few images per year of a given area. For example, IGN campaigns, which are the reference for aerial surveys in France, are performed every 3-5 years. The evolution of remote sensing technologies provides the opportunity to enlarge studies using time series in several ways.

First, data acquired from new sensors, such as LiDAR and hyperspectral sensors, are becoming available as time series. They provide the opportunity to monitor changes in specific features of riparian vegetation, such as canopy height and species composition. For example, a LiDAR survey covers the entire region of Wallonia (southern Belgium) every six years. In France, in the framework of the Litto3D program, ca. 45,000 km² of coast (bathymetry included) will be covered with a dense LiDAR survey (2-15 ground points/m² for emerged areas). This cover will be regularly updated to monitor sediment dynamics and erosion processes.

In addition, acquisition frequency has increased. For example, UAVs can acquire dense time series easily. High-resolution satellite images such as Sentinel-1 and Sentinel-2 (four bands at 10 m resolution) provide images of the Earth's entire surface every few days. More recently, CubeSat constellations provide higher resolution and higher frequency. For example, the Dove constellation (Planet Labs, Inc., San Francisco, CA, USA) provides resolution up to 3 m and daily coverage. This increased frequency of image acquisition provides new opportunities to study rapid riparian vegetation processes, including intra-annual ones such as phenology and impacts of flood events. From a more operational viewpoint, it could be used to obtain information about riparian vegetation over large areas following floods to locate areas to be managed.

Finally, the increased availability of dense time series increases the potential of remote sensing data to provide a picture of riparian vegetation at a given time. Time series from satellite images can be used to describe vegetation by using phenology or management operations to distinguish species. For example, Rapinel et al. (2019) used Sentinel-2 time series to classify grassland plant communities in a temperate floodplain using the relationship between inundation, grassland management and vegetation composition. Similarly, Michez et al. (2016b) used UAV time series to distinguish riparian tree species using images acquired during several phenological stages (from spring to fall).

4.3. SCALING UP CHARACTERIZATION OF RIPARIAN VEGETATION

Our results highlight that the features of riparian vegetation are strongly correlated with the extent of studies. For example, physiological processes such as evapotranspiration and health status are extracted mainly from low- or medium-resolution multispectral data (MODIS and Landsat) over large study areas. Conversely, features related to species composition are mapped mainly from high-resolution data (< 10 m), such as high-resolution satellite multispectral images (17% of studies that mapped species composition), UAV or aerial orthophotos (> 50% of studies), or LiDAR data (20% of studies). We showed that upscaling such data was challenging beyond a few dozen km of river. However, at this scale, remote sensing would be a particularly useful alternative to field campaigns or photointerpretation. Species classification methods that are more robust to upscaling still need to be developed, as Fassnacht et al. (2016) indicated in a literature review of classification of forest tree species using remote sensing. Since they are less subject to changing atmospheric conditions, 3D data can simplify development of indicators that are more stable across a large study area. Ultimately, 3D approaches will be promoted by the growing availability of nationwide systematic LiDAR surveys across countries.

5. PERSPECTIVES FOR RIPARIAN VEGETATION MANAGEMENT

Remote sensing scientists often develop tools to support management of riparian vegetation. Overall, 38% of the abstracts in our database contained the words “management”, “restoration” or their derivatives. However, the contribution of remote sensing tools to actual management operations was difficult to evaluate from the review, since scientists usually do not describe how river managers use the tools or knowledge they have developed.

Therefore, we used an expert-based approach to analyze this issue. We provide examples of operational or near-operational applications to illustrate the potential use of remote sensing technologies to support riparian vegetation management. We also discuss how tools and knowledge that scientists produce can be transferred to river managers.

5.1. EXAMPLES OF NEAR-OPERATIONAL APPLICATIONS

To illustrate the perspectives of applications for riparian managers, we chose three contrasting examples that are particularly relevant for the riparian context: eradication of invasive plant species, monitoring ecological integrity at the regional scale and maintenance of hydraulic conveyance.

5.1.1. MANAGING INVASIVE PLANT SPECIES AT THE LOCAL SCALE

Riparian managers often conduct programs to eradicate invasive plant species. These programs require identifying and locating individuals prior to eradication measures and subsequent monitoring of invasive cover (i.e. to ensure that practices were effective and that the species do not re-emerge) (Vaz et al., 2018). These actions can be performed with UAVs that combine high spatial resolution (useful for detecting invasive plant species at an early stage, before they form large clumps) and high temporal resolution (invasive plant species are often more distinct from the background during a particular phenological phase, according to Manfreda et al., 2018). Many studies have shown that detecting invasive plant species using a UAV could outperform ground surveys in terms of cost, effectiveness and risk mitigation for operators (Martin et al., 2018; Michez et al., 2016a). In the future, real-time or onboard processing (i.e. analysis of streamed imagery) will enable detection and eradication steps to be performed at the same time (Hill and Babbar-Sebens, 2019).

However, several challenges must be addressed before using UAVs in eradication operations. Legal restrictions may apply for flight outside the visual line of sight or for overflight of urban areas. Technical challenges must also be addressed, in particular the autonomy of UAVs. The detection period (e.g. during a particular phenological phase) must be adequate for eradication requirements. Perhaps most important, river managers must have access to skilled staff who are able to pilot the UAV and process the images based on the needs of

riparian managers. The staff can be recruited and trained within the organization, or work for an exterior contracting organization. For invasive species, work is often concentrated in time, and skilled personnel must be available at that time.

5.1.2. MONITORING ECOLOGICAL INTEGRITY AT THE REGIONAL SCALE

Managers of riparian vegetation at the regional or national scale sometimes need information about the entire river network to assess effects of policies or define management strategies (e.g. to prioritize which zones should be restored). For example, all EU member states must monitor the state of riparian ecosystems to comply with the Water Framework Directive (WFD), which promotes a good health status of European rivers. These assessments have historically been performed during field visits to sites sampled throughout each river network (Hering et al., 2010; Munné et al., 2003). They can include remote sensing techniques in different ways. We briefly present two contrasting approaches to include remote sensing in ecological assessments: a sampling- and photointerpretation-based approach using aerial images, or the use of regional LiDAR data to map riparian structural attributes automatically.

In the first approach, aerial images can be integrated with minor adaptations into a traditional field-based, sampling approach. Aerial images are used to target sampling sites (e.g. where riparian vegetation is present) and to perform certain aspects of the assessment, especially those that require less specific information at a larger scale. For example, the Riparian Quality Index, initially developed for Iberian rivers, includes measurements of width, continuity, strata, composition, regeneration, bank condition, lateral connectivity and substratum (González del Tánago and García de Jalón, 2011). Width, continuity and strata can be described using aerial imagery, while other attributes are assessed in the field. More attributes can be mapped with remote sensing using other sources of data (e.g. 2 cm aerial images in Booth et al., 2007), but traditional aerial images have the advantage of being widely, and often freely, available.

In the second approach, regional 3D data can be used to assess riparian features in a spatially continuous manner. Riparian attributes are calculated with a high level of automation and can be updated at the same frequency as that of the LiDAR cover. For example, [Michez et al. \(2017\)](#) used this approach using LiDAR and photogrammetric point clouds to map riparian buffer attributes along 12,000 km of rivers (vegetation continuity, height and overhang; channel width and sinuosity; and lateral connectivity (indicated by emerged channel depth)). In this case, the strength of 3D data is that the 3D component is homogenous at the regional scale (unlike spectral data) and can extract attributes of the channel even when it is hidden by vegetation.

5.1.3. IMPROVING FLOOD MODELING WITH BETTER ESTIMATES OF FLOODPLAIN ROUGHNESS

Many regions of the world must address significant and increasing threats of flooding, as well as the need to conserve riparian ecosystems (Straatsma et al., 2019). Floodplain vegetation can influence flood risk by increasing hydraulic roughness. In the Netherlands, where these challenges are particularly acute, several remote sensing applications integrate riparian vegetation management more into flood mitigation strategies.

One example includes a legal map produced to describe the maximum roughness of vegetation cover allowed within the floodplains of major Dutch rivers. The legal map uses a historical situation as a target reference (Rijkswaterstaat, 2014). To support use of this legal map, Deltares (an independent applied research institute) and the Rijkswaterstaat (the administration responsible for river management) developed an online vegetation-mapping tool. In the Google Earth Engine environment, users can easily classify the vegetation cover observed on recent Sentinel-2 images to ensure that it complies with the legal standard. The tool is available on smartphones and can be used in the field. Vegetation can be compared to the map before each winter, when most floods occur. The tool provides information about the areas on which management practices should focus, following a dialogue with the landowners concerned (Penning, 2018).

Modeling approaches are also useful to support decisions. To prevent flood damage in Dutch deltas, multiple practices, such as raising dikes or removing riparian vegetation, must be implemented in a coordinated

manner. Straatsma and Kleinhans (2018) developed the RiverScape toolbox, which can model effects of riparian cutting scenarios on flow using hydrological and spatial data (including a DTM, vegetation map and its associated roughness coefficients). The toolbox can optimize the location of cutting operations to reduce water levels during floods.

5.2. INTERACTION BETWEEN TOOL DEVELOPMENT AND RIPARIAN MANAGEMENT

This review has addressed the potential of remote sensing tools to support research and management of riparian vegetation. However, scientific publications do not usually directly address the transfer of these tools from researchers to managers and their implementation in real contexts.

We distinguish three main steps in this process (Figure 10). First, managers and remote sensing experts must work together to define clear **objectives**. Second, the **development** step implies a technological phase. Third, thorough **assessment** must be performed for accuracy, reliability and relevance for managers. **Critical thinking** is required throughout this process because the choice of a remote sensing approach is not neutral and has implications for how riparian vegetation is managed. Scientific articles often describe the development step and assessment of accuracy well but rarely discuss the reasons and interactions that led scientists and managers to choose remote sensing tools for their operations.

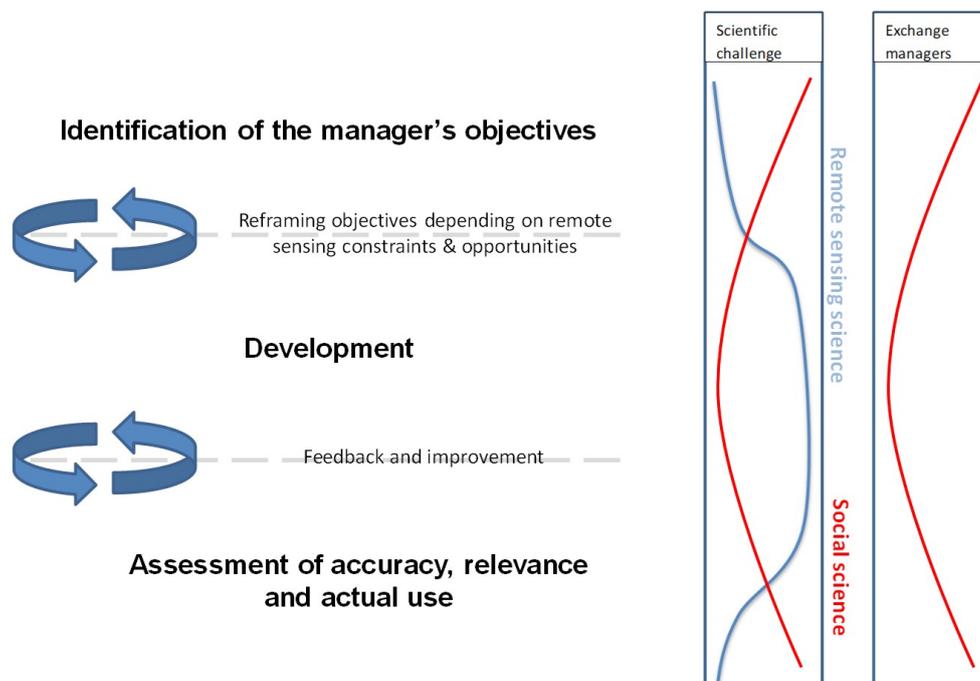


FIGURE 10. CONCEPTUAL FRAMEWORK OF THE TRANSFER OF REMOTE SENSING TOOLS FROM SCIENTISTS TO MANAGERS. ON THE GRAPHICS TO THE RIGHT, THE HORIZONTAL AXES REPRESENT SCIENTIFIC CHALLENGE AND EXCHANGE DEGREE TO BE PLANNED BETWEEN MANAGERS AND RESEARCHERS (FROM LOW TO HIGH), WHILE THE VERTICAL AXES REPRESENT THEIR DYNAMICS FROM START TO FINISH.

5.2.1. IDENTIFYING THE ISSUES/NEEDS OF RIPARIAN VEGETATION MANAGERS

Riparian vegetation managers can use remote sensing approaches for different reasons. First, they may encounter a recurring situation for which no satisfying solution can be found (e.g. invasive plant species). Second, they may use remote sensing approaches to implement new legislation or new considerations that require new methods (e.g. additional landscape analyses for WFD monitoring). Finally, the use of remote sensing can also be promoted by a specific opportunity (e.g. new processing techniques and or sensors) rather than an operational need.

The first step in implementing a remotely sensed application is to define the needs and objectives of riparian vegetation managers. Key issues must be addressed, such as the features to be mapped, the scale of observation, the time required to obtain usable information and the frequency of updating. Objectives can be refined during the development step, depending on the tradeoffs between costs and image quality. Nevertheless, thoroughly defining the objectives beforehand is clearly a factor of success (Kennedy et al., 2009).

However, thorough definition of objectives is not straightforward. To translate monitoring objectives into a remote sensing approach requires an explicit space for collaboration between remote sensing specialists and managers (Kennedy et al., 2009). Managers are often unsure about the operational potential of remote sensing data and techniques (Vanden Borre et al., 2011). This is increasingly true, since new technologies (e.g. satellites, UAVs) seem to be developed very quickly, and even faster than the applications for using them. Therefore, realistic monitoring objectives must be defined along with remote sensing specialists. Moreover, field approaches and remote sensing approaches often are not perfectly interchangeable (Dufour et al., 2012). Challenging the work routine of managers might be required to fully benefit from remote sensing approaches. The collaborative process should thus be open enough to consider adapting work routines. Similarly, when relevant, managers and scientists from different fields must be involved. It is important to combine a variety of scientific perspectives (e.g. geomatic, landscape planning, riparian ecology) to avoid too narrow or inappropriate solutions.

In many cases at this stage, riparian vegetation is not the center of management operations. Many studies and management operations focus on the river channel and its hydrological and geomorphological components (e.g. WFD monitoring, floodplain roughness modeling).

5.2.2. DEVELOPING APPLICATIONS THAT USE REMOTE SENSING DATA

Once the objectives have been clearly identified, the next step is to develop the solution to use remote sensing data to pursue the manager's objectives. Several stakeholders are involved in this process. We artificially distinguish data "producers" from the "developers".

We consider "producers" the stakeholders who provide rough datasets, such as raw satellite images or raw ancillary data (e.g. national space agencies such as NASA and CNES, UAV constructors). While they do not interact closely with riparian vegetation managers, their role is important in the long run since they set the agenda for the main future developments of new remote sensing technologies. More directly, they can promote the use of remote sensing data for natural resource managers by making the data affordable and easier to use, as mentioned (section 4.1.).

We consider "developers" the stakeholders who develop tools that use raw remote sensing data. They may interact more closely with riparian vegetation managers and provide solutions that are tailored to the latter's needs through the previously mentioned space for collaboration. The main stakeholders in this category are academic and research institutes, as well as commercial or non-academic organizations, that use remote sensing data. In theory, the needs identified define the type of stakeholders involved. For example, if the manager's issue has scientific relevance (e.g. understanding the spread of an invasive species not studied before), academics would logically be involved. If no scientific issue is identified, however, then commercial or non-academic organizations are more appropriate.

The fixed costs of implementing a remote sensing approach seem relatively high and can be prohibitive for many local managers, even though free solutions increasingly appear on the market. These costs include designing the method, deploying the platform or acquiring the minimum number of satellite images and possibly training personnel. Moreover, performing certain analyses requires technical skills (e.g. object-based image analysis, machine learning approaches, LiDAR full waveform analysis). Therefore, remote sensing could have greater relevance when the area to be mapped is large and/or the operation must be repeated several

times (Johansen et al., 2007). However, many stakeholders with different objectives are involved, since riparian vegetation covers large geographical areas. This can reduce the potential for economies of scale, whether for river managers trying to develop their own expertise or for businesses offering their services. This narrow market provides relatively limited opportunities for companies to develop specific tools adapted for this vegetation type. Indeed, we do not expect specific UAV applications to become as developed for riparian vegetation as they are for precision agriculture.

To address the challenge of attaining “critical mass” for riparian vegetation, we suggest a more collaborative approach, as described by Steiniger and Hay (2009). Processing routines developed by remote sensing scientists could be embedded into OpenAccess toolboxes. To benefit a large audience, these tools must be robust by having little sensitivity to situations that differ slightly from those for which they were created. For managers to use them, they need to be flexible and integrate easily with other processing routines or platforms (e.g. GIS platforms) (Vanden Borre et al., 2011). Finally, they should be based on widely available data. These tools could be collected in community repositories along with other tools for river or ecosystem management, along with with freely available datasets, as suggested by Tomsett and Leyland (2019). These tools could form a foundation that commercial companies, researchers and managers could adapt to specific projects.

5.2.3. ASSESSMENT AND FEEDBACK

The final step in applying remote sensing to riparian management involves accurate and effective assessment of the maps produced and the potential for future monitoring. Accuracy involves the statistical validity of the product, which is the conformity of the map to reference data (e.g. thematic accuracy, in the case of classification). This step is crucial because it indicates the extent to which the map can be trusted. Remote sensing specialists usually consider it a central element, although controversy remains on the reliability of popular accuracy assessment methods (Pontius and Millones, 2011). Moreover, users must be cautious when reproducing the method at another site, since accuracy is often assessed for small test sites, and robustness is often not assessed sufficiently (Fassnacht et al., 2016).

However, the relevance of a remote sensing approach cannot be reduced to its accuracy. The relevance of the information for management purposes must consider the costs and benefits of obtaining such information (Kennedy et al., 2009). We argue that temporality should be considered when addressing this aspect. The true effectiveness of a tool is often observed long after it is first produced. Moreover, the issue of using remote sensing data in future monitoring (or not) must be considered. Riparian vegetation often requires monitoring for several years. For example, after a restoration action, vegetation must be monitored in the short term (i.e. after one year) and the long term (i.e. after 5-10 years). Consequently, it is important to define which stakeholders are involved in this future monitoring (the initial producer of the map, the managers themselves or an external stakeholder) and which methods will be used. Because remote sensing data continually change, a method adapted at one time (e.g. one year after a restoration action) may not be as adapted several years later, or more powerful approaches may then be available.

The ease of use the tools developed and their integration into existing workflows are also central aspects in whether a manager will adopt remote sensing tools (Vanden Borre et al., 2011). Ultimately, what matters for managers is the degree of “satisfaction” and their real use of the tool. We argue that it is crucial to obtain feedback from managers about the real use of the maps and features produced using remote sensing data. This feedback would help to develop tools that are more adapted to the managers’ needs.

5.2.4. ISSUES BEYOND THE REMOTE SENSING DISCIPLINE

The development and use of remote sensing tools to manage riparian vegetation is not only a technical issue. It raises at least three particular issues that must be addressed in an interdisciplinary or even transdisciplinary manner. First, the information must be scientifically relevant from a thematic perspective. For example, LiDAR data make it possible to measure vegetation density within a given range of elevation (e.g. 4-10 m). However, whether this information is sufficient or relevant to assess a particular function of riparian vegetation must be

discussed with experts from different disciplines (e.g. ecologists, hydrologists). Second, understanding the needs is crucial, as mentioned (section 5.2.1). A co-construction approach that includes managers and social scientists must be adopted as soon as possible to avoid purely data-driven processes. Third, critical feedback about the use of remote sensing tools is also needed afterwards. Using these tools to assess environmental patterns and processes or to map natural resources is clearly not neutral. In some cases, these methods exclude certain stakeholders who do not have access to the technology, limit the understanding of certain complex phenomena and generate controversial data (e.g. Fairhead and Leach (1998); Harwell (2000); Turner and Taylor (2003); Rajão (2013)). Therefore, widespread use of these tools must be combined with critical understanding of sociological and cultural effects, along with the use of additional approaches to counterbalance potential negative effects. Social scientists should be involved throughout the process to address these issues.

6. CONCLUSION

We found a substantial body of literature in which remote sensing approaches were used to study riparian vegetation. The approaches became considerably popular at the turn of the millennium, but their relative use in riparian vegetation studies remains limited (ca. 4%). Development of new sensors and platforms has improved remote sensing approaches. LiDAR data have increased the ability to characterize vegetation structure, and high-resolution multispectral data hold the promise of mapping species composition in species-rich riparian forests. Riparian ecosystems are highly dynamic, and the multi-temporal nature of riparian remote sensing studies is central. To date, diachronic analyses have relied essentially on aerial photographs. The increasing availability of other data time series will promote the study of complex and subtle phenomena, beyond changes in the extent of riparian forests or plant succession. Most studies focus on the local-to-river segment scale. Large-scale studies are based on medium-resolution data. Algorithms are needed to process high-resolution data that is robust to upscaling. Spectral heterogeneity makes upscaling the study of species composition using spectral data more challenging than upscaling the study of vegetation structure using 3D data.

It is often suggested that remote sensing approaches can contribute to management of riparian vegetation by providing objective, continuous and up-to-date data for a large area. However, this contribution was difficult to determine via a review of the scientific literature. An extensive review of the gray literature could provide further insight into this subject. Our research suggests that the potential for applying remote sensing to riparian vegetation management is increasing, due to innovation in platforms and sensors. We suggest that a collaborative effort is required to make remote sensing approaches more robust and available, both in terms of cost and ease of use. This effort would also promote the study of riparian vegetation in areas where it is less studied. However, implementing a remote sensing approach in management operations requires developing a space for collaboration between remote sensing specialists and riparian managers.

7. REFERENCES

- Akasaka, T., Akasaka, M., Yanagawa, H., 2010. Relative importance of the environmental factors at site and landscape scales for bats along the riparian zone. *Landsc. Ecol. Eng.* 6, 247–255. <https://doi.org/10.1007/s11355-010-0105-4>
- Alaibakhsh, M., Emelyanova, I., Barron, O., Sims, N., Khiadani, M., Mohyeddin, A., 2017. Delineation of riparian vegetation from Landsat multi-temporal imagery using PCA. *Hydrol. Process.* 31, 800–810. <https://doi.org/10.1002/hyp.11054>
- Anderson, K., Gaston, K.J., 2013. Lightweight unmanned aerial vehicles will revolutionize spatial ecology. *Front. Ecol. Environ.* 11, 138–146. <https://doi.org/10.1890/120150>
- Antonarakis, A.S., Richards, K.S., Brasington, J., 2008. Object-based land cover classification using airborne LiDAR. *Remote Sens. Environ.* 112, 2988–2998. <https://doi.org/10.1016/j.rse.2008.02.004>

- Arroyo, L.A., Johansen, K., Armston, J., Phinn, S., 2010. Integration of LiDAR and QuickBird imagery for mapping riparian biophysical parameters and land cover types in Australian tropical savannas. *For. Ecol. Manag.* 259, 598–606. <https://doi.org/10.1016/j.foreco.2009.11.018>
- Ashraf, S., Brabyn, L., Hicks, B.J., Collier, K., 2010. Satellite remote sensing for mapping vegetation in New Zealand freshwater environments: A review. *N. Z. Geogr.* 66, 33–43. <https://doi.org/10.1111/j.1745-7939.2010.01168.x>
- Beechie, T., Pess, G., Roni, P., Giannico, G., 2008. Setting river restoration priorities: a review of approaches and a general protocol for identifying and prioritizing actions. *North Am. J. Fish. Manag.* 28, 891–905.
- Belletti, B., Dufour, S., Piégay, H., 2015a. What is the Relative Effect of Space and Time to Explain the Braided River Width and Island Patterns at a Regional Scale? *River Res. Appl.* 31, 1–15. <https://doi.org/10.1002/rra.2714>
- Belletti, B., Rinaldi, M., Buijse, A.D., Gurnell, A.M., Mosselman, E., 2015b. A review of assessment methods for river hydromorphology. *Environ. Earth Sci.* 73, 2079–2100. <https://doi.org/10.1007/s12665-014-3558-1>
- Belluco, E., Camuffo, M., Ferrari, S., Modenese, L., Silvestri, S., Marani, A., Marani, M., 2006. Mapping salt-marsh vegetation by multispectral and hyperspectral remote sensing. *Remote Sens. Environ.* 105, 54–67. <https://doi.org/10.1016/j.rse.2006.06.006>
- Bendix, J., Stella, J.C., 2013. Riparian Vegetation and the Fluvial Environment: A Biogeographic Perspective. *Treatise Geomorphol.* 53–74. <https://doi.org/10.1016/B978-0-12-374739-6.00322-5>
- Bertoldi, W., Drake, N.A., Gurnell, A.M., 2011. Interactions between river flows and colonizing vegetation on a braided river: exploring spatial and temporal dynamics in riparian vegetation cover using satellite data. *Earth Surf. Process. Landf.* 36, 1474–1486. <https://doi.org/10.1002/esp.2166>
- Booth, D.T., Cox, S.E., Simonds, G., 2007. Riparian monitoring using 2-cm GSD aerial photography. *Ecol. Indic.* 7, 636–648. <https://doi.org/10.1016/j.ecolind.2006.07.005>
- Bowler, D.E., Mant, R., Orr, H., Hannah, D.M., Pullin, A.S., 2012. What are the effects of wooded riparian zones on stream temperature? *Environ. Evid.* 1, 3. <https://doi.org/10.1186/2047-2382-1-3>
- Bucha, T., Slávik, M., 2013. Improved methods of classification of multispectral aerial photographs: Evaluation of floodplain forests in the inundation area of the Danube. *Folia For. Pol. Ser. - For.* 55. <https://doi.org/10.2478/ffp-2013-0007>
- Bunn, S.E., Abal, E.G., Smith, M.J., Choy, S.C., Fellows, C.S., Harch, B.D., Kennard, M.J., Sheldon, F., 2010. Integration of science and monitoring of river ecosystem health to guide investments in catchment protection and rehabilitation. *Freshw. Biol.* 55, 223–240. <https://doi.org/10.1111/j.1365-2427.2009.02375.x>
- Cadol, D., Wine, M.L., 2017. Geomorphology as a first order control on the connectivity of riparian ecohydrology. *Geomorphology, Connectivity in Geomorphology from Binghamton 2016* 277, 154–170. <https://doi.org/10.1016/j.geomorph.2016.06.022>
- Caissie, D., 2006. The thermal regime of rivers: a review. *Freshw. Biol.* 51, 1389–1406. <https://doi.org/10.1111/j.1365-2427.2006.01597.x>
- Capon, S.J., Pettit, N.E., 2018. Turquoise is the new green: Restoring and enhancing riparian function in the Anthropocene. *Ecol. Manag. Restor.* 19, 44–53. <https://doi.org/10.1111/emr.12326>
- Carbonneau, P., Piégay, H., 2012. Future prospects and challenges for river scientists and managers, in: *Fluvial Remote Sensing for Science and Management*. John Wiley & Sons, Ltd. <https://doi.org/10.1002/9781119940791>
- Carli, C.M., Bayley, S.E., 2015. River connectivity and road crossing effects on floodplain vegetation of the upper Columbia River, Canada. *Écoscience* 22, 97–107. <https://doi.org/10.1080/11956860.2015.1121705>
- Claggett, P.R., Okay, J.A., Stehman, S.V., 2010. Monitoring Regional Riparian Forest Cover Change Using Stratified Sampling and Multiresolution Imagery1. *JAWRA J. Am. Water Resour. Assoc.* 46, 334–343. <https://doi.org/10.1111/j.1752-1688.2010.00424.x>
- Congalton, R.G., Birch, K., Jones, R., Schriever, J., 2002. Evaluating remotely sensed techniques for mapping riparian vegetation. *Comput. Electron. Agric.* 37, 113–126. [https://doi.org/10.1016/S0168-1699\(02\)00108-4](https://doi.org/10.1016/S0168-1699(02)00108-4)
- Cunningham, S.C., Griffioen, P., White, M.D., Nally, R.M., 2018. Assessment of ecosystems: A system for rigorous and rapid mapping of floodplain forest condition for Australia’s most important river. *Land Degrad. Dev.* 29, 127–137. <https://doi.org/10.1002/ldr.2845>
- Curran, J.C., Hession, W.C., 2013. Vegetative impacts on hydraulics and sediment processes across the fluvial system. *J. Hydrol.* 505, 364–376. <https://doi.org/10.1016/j.jhydrol.2013.10.013>

- Dahm, C.N., Cleverly, J.R., Coonrod, J.E.A., Thibault, J.R., McDonnell, D.E., Gilroy, D.J., 2002. Evapotranspiration at the land/water interface in a semi-arid drainage basin. *Freshw. Biol.* 47, 831–843. <https://doi.org/10.1046/j.1365-2427.2002.00917.x>
- Dauwalter, D.C., Fesenmyer, K.A., Bjork, R., 2015. Using Aerial Imagery to Characterize Redband Trout Habitat in a Remote Desert Landscape. *Trans. Am. Fish. Soc.* 144, 1322–1339. <https://doi.org/10.1080/00028487.2015.1088471>
- Doody, T.M., Lewis, M., Benyon, R.G., Byrne, G., 2014. A method to map riparian exotic vegetation (*Salix* spp.) area to inform water resource management. *Hydrol. Process.* 28, 3809–3823. <https://doi.org/10.1002/hyp.9916>
- Dufour, S., Bernez, I., Betbeder, J., Corgne, S., Hubert-Moy, L., Nabucet, J., Rapinel, S., Sawtschuk, J., Trollé, C., 2013. Monitoring restored riparian vegetation: how can recent developments in remote sensing sciences help? *Knowl. Manag. Aquat. Ecosyst.* 10. <https://doi.org/10.1051/kmae/2013068>
- Dufour, S., Muller, E., Straatsma, M., Corgne, S., 2012. Image Utilisation for the Study and Management of Riparian Vegetation: Overview and Applications, in: *Fluvial Remote Sensing for Science and Management*. John Wiley & Sons, Ltd. <https://doi.org/10.1002/9781119940791>
- Dufour, S., Piégay, H., 2009. From the myth of a lost paradise to targeted river restoration: forget natural references and focus on human benefits. *River Res. Appl.* 25, 568–581. <https://doi.org/10.1002/rra.1239>
- Dufour, S., Rinaldi, M., Piégay, H., Michalon, A., 2015. How do river dynamics and human influences affect the landscape pattern of fluvial corridors? Lessons from the Magra River, Central–Northern Italy. *Landsc. Urban Plan.* 134, 107–118. <https://doi.org/10.1016/j.landurbplan.2014.10.007>
- Dufour, S., Rodríguez-González, P.M., Laslier, M., 2019. Tracing the scientific trajectory of riparian vegetation studies: Main topics, approaches and needs in a globally changing world. *Sci. Total Environ.* 653, 1168–1185. <https://doi.org/10.1016/j.scitotenv.2018.10.383>
- Dunford, R., Michel, K., Gagnage, M., Piégay, H., Trémelo, M.-L., 2009. Potential and constraints of Unmanned Aerial Vehicle technology for the characterization of Mediterranean riparian forest. *Int. J. Remote Sens.* 30, 4915–4935. <https://doi.org/10.1080/01431160903023025>
- Dutta, D., Wang, K., Lee, E., Goodwell, A., Woo, D.K., Wagner, D., Kumar, P., 2017. Characterizing Vegetation Canopy Structure Using Airborne Remote Sensing Data. *IEEE Trans. Geosci. Remote Sens.* 55, 1160–1178. <https://doi.org/10.1109/TGRS.2016.2620478>
- Fairhead, J., Leach, M., 1998. Reconsidering the extent of deforestation in twentieth century West Africa. *Unasylva* 49/192 38–46.
- Fassnacht, F.E., Latifi, H., Stereńczak, K., Modzelewska, A., Lefsky, M., Waser, L.T., Straub, C., Ghosh, A., 2016. Review of studies on tree species classification from remotely sensed data. *Remote Sens. Environ.* 186, 64–87. <https://doi.org/10.1016/j.rse.2016.08.013>
- Fausch, K.D., Torgersen, C.E., Baxter, C.V., Li, H.W., 2002. Landscapes to Riverscapes: Bridging the Gap between Research and Conservation of Stream Fishes. *BioScience* 52, 483–498. [https://doi.org/10.1641/0006-3568\(2002\)052\[0483:LTRBTG\]2.0.CO;2](https://doi.org/10.1641/0006-3568(2002)052[0483:LTRBTG]2.0.CO;2)
- Fédération des Conservatoires d’espaces naturels, 2018. Les nouvelles technologies dans l’étude des milieux humides : exemples d’applications.
- Fernandes, M.R., Aguiar, F.C., Ferreira, M.T., Pereira, J.M.C., 2013a. Spectral separability of riparian forests from small and medium-sized rivers across a latitudinal gradient using multispectral imagery. *Int. J. Remote Sens.* 34, 2375–2401. <https://doi.org/10.1080/01431161.2012.744491>
- Fernandes, M.R., Aguiar, F.C., Silva, J.M.N., Ferreira, M.T., Pereira, J.M.C., 2014. Optimal attributes for the object based detection of giant reed in riparian habitats: A comparative study between Airborne High Spatial Resolution and WorldView-2 imagery. *Int. J. Appl. Earth Obs. Geoinformation* 32, 79–91. <https://doi.org/10.1016/j.jag.2014.03.026>
- Fernandes, M.R., Aguiar, F.C., Silva, J.M.N., Ferreira, M.T., Pereira, J.M.C., 2013b. Spectral discrimination of giant reed (*Arundo donax* L.): A seasonal study in riparian areas. *ISPRS J. Photogramm. Remote Sens.* 80, 80–90. <https://doi.org/10.1016/j.isprsjprs.2013.03.007>
- Forzieri, G., Castelli, F., Preti, F., 2012. Advances in remote sensing of hydraulic roughness. *Int. J. Remote Sens.* 33, 630–654. <https://doi.org/10.1080/01431161.2010.531788>
- Forzieri, G., Tanteri, L., Moser, G., Catani, F., 2013. Mapping natural and urban environments using airborne multi-sensor ADS40–MIVIS–LiDAR synergies. *Int. J. Appl. Earth Obs. Geoinformation* 23, 313–323. <https://doi.org/10.1016/j.jag.2012.10.004>

- Fullerton, A.H., Beechie, T.J., Baker, S.E., Hall, J.E., Barnas, K.A., 2006. Regional patterns of riparian characteristics in the interior Columbia River basin, Northwestern USA: applications for restoration planning. *Landsc. Ecol.* 21, 1347–1360. <https://doi.org/10.1007/s10980-006-0017-8>
- Georges, B., Brostaux, Y., Claessens, H., Degré, A., Huylensbroeck, L., Lejeune, P., Piégay, H., Michez, A., In press. Can water level stations be used for thermal assessment in aquatic ecosystem? *River Res. Appl.* n/a. <https://doi.org/10.1002/rra.3520>
- Gergel, S.E., Stange, Y., Coops, N.C., Johansen, K., Kirby, K.R., 2007. What is the Value of a Good Map? An Example Using High Spatial Resolution Imagery to Aid Riparian Restoration. *Ecosystems* 10, 688–702. <https://doi.org/10.1007/s10021-007-9040-0>
- Goetz, S.J., 2006. Remote Sensing of Riparian Buffers: Past Progress and Future Prospects. *JAWRA J. Am. Water Resour. Assoc.* 42, 133–143. <https://doi.org/10.1111/j.1752-1688.2006.tb03829.x>
- Goetz, W.E., 2002. Developing a predictive model for identifying riparian communities at an ecoregion scale in Idaho and Wyoming /.
- Gong, P., Wang, J., Yu, L., Zhao, Yongchao, Zhao, Yuanyuan, Liang, L., Niu, Z., Huang, X., Fu, H., Liu, S., Li, C., Li, X., Fu, W., Liu, C., Xu, Y., Wang, X., Cheng, Q., Hu, L., Yao, W., Zhang, Han, Zhu, P., Zhao, Z., Zhang, Haiying, Zheng, Y., Ji, L., Zhang, Y., Chen, H., Yan, A., Guo, J., Yu, Liang, Wang, L., Liu, X., Shi, T., Zhu, M., Chen, Y., Yang, G., Tang, P., Xu, B., Giri, C., Clinton, N., Zhu, Z., Chen, Jin, Chen, Jun, 2013. Finer resolution observation and monitoring of global land cover: first mapping results with Landsat TM and ETM+ data. *Int. J. Remote Sens.* 34, 2607–2654. <https://doi.org/10.1080/01431161.2012.748992>
- Gong, Z., Cui, T., Pu, R., Lin, C., Chen, Y., 2015. Dynamic simulation of vegetation abundance in a reservoir riparian zone using a sub-pixel Markov model. *Int. J. Appl. Earth Obs. Geoinformation* 35, 175–186. <https://doi.org/10.1016/j.jag.2014.09.004>
- González del Tánago, M., Bejarano, M.D., García de Jalón, D., Schmidt, J.C., 2015. Biogeomorphic responses to flow regulation and fine sediment supply in Mediterranean streams (the Guadalete River, southern Spain). *J. Hydrol.* 528, 751–762. <https://doi.org/10.1016/j.jhydrol.2015.06.065>
- González del Tánago, M., García de Jalón, D., 2011. Riparian Quality Index (RQI): a methodology for characterising and assessing the environmental conditions of riparian zones. *Limnética* 30, 0235–0254.
- González, E., González-Sanchis, M., Cabezas, Á., Comín, F.A., Muller, E., 2010. Recent Changes in the Riparian Forest of a Large Regulated Mediterranean River: Implications for Management. *Environ. Manage.* 45, 669–681. <https://doi.org/10.1007/s00267-010-9441-2>
- González, E., Sher, A.A., Tabacchi, E., Masip, A., Poulin, M., 2015. Restoration of riparian vegetation: A global review of implementation and evaluation approaches in the international, peer-reviewed literature. *J. Environ. Manage.* 158, 85–94. <https://doi.org/10.1016/j.jenvman.2015.04.033>
- Groeneveld, D.P., Watson, R.P., 2008. Near-infrared discrimination of leafless saltcedar in wintertime Landsat TM. *Int. J. Remote Sens.* 29, 3577–3588. <https://doi.org/10.1080/01431160701711078>
- Hamandawana, H., Chanda, R., 2013. Environmental change in and around the Okavango Delta during the nineteenth and twentieth centuries. *Reg. Environ. Change* 13, 681–694. <https://doi.org/10.1007/s10113-012-0367-5>
- Harwell, E., 2000. Remote Sensibilities: Discourses of Technology and the Making of Indonesia's Natural Disaster. *Dev. Change* 31, 307–340. <https://doi.org/10.1111/1467-7660.00156>
- Hering, D., Borja, A., Carstensen, J., Carvalho, L., Elliott, M., Feld, C.K., Heiskanen, A.-S., Johnson, R.K., Moe, J., Pont, D., Solheim, A.L., de Bund, W. van, 2010. The European Water Framework Directive at the age of 10: A critical review of the achievements with recommendations for the future. *Sci. Total Environ.* 408, 4007–4019. <https://doi.org/10.1016/j.scitotenv.2010.05.031>
- Hill, D.J., Babbar-Sebens, M., 2019. Promise of UAV-Assisted Adaptive Management of Water Resources Systems. *J. Water Resour. Plan. Manag.* 145, 02519001. [https://doi.org/10.1061/\(ASCE\)WR.1943-5452.0001081](https://doi.org/10.1061/(ASCE)WR.1943-5452.0001081)
- Hupp, C.R., 1992. Riparian Vegetation Recovery Patterns Following Stream Channelization: A Geomorphic Perspective. *Ecology* 73, 1209–1226. <https://doi.org/10.2307/1940670>
- Jalonen, J., Järvelä, J., Virtanen, J.-P., Vaaja, M., Kurkela, M., Hyyppä, H., 2015. Determining Characteristic Vegetation Areas by Terrestrial Laser Scanning for Floodplain Flow Modeling. *Water* 7, 420–437. <https://doi.org/10.3390/w7020420>
- Jansen, B.J.M., Backx, J.J.G.M., 1998. Biologische monitoring zoete rijkswateren : ecotopenkartering Rijntakken-oost 1997, RIZA rapport;98.054. RIZA, Lelystad.
- Johansen, K., Phinn, S., Dixon, I., Douglas, M., Lowry, J., 2007. Comparison of image and rapid field assessments of riparian zone condition in Australian tropical savannas. *For. Ecol. Manag.* 240, 42–60. <https://doi.org/10.1016/j.foreco.2006.12.015>

- Johansen, K., Phinn, S., Witte, C., 2010. Mapping of riparian zone attributes using discrete return LiDAR, QuickBird and SPOT-5 imagery: Assessing accuracy and costs. *Remote Sens. Environ.* 114, 2679–2691. <https://doi.org/10.1016/j.rse.2010.06.004>
- Kalliola, R., Puhakka, M., 1988. River Dynamics and Vegetation Mosaicism: A Case Study of the River Kamajohka, Northernmost Finland. *J. Biogeogr.* 15, 703–719. <https://doi.org/10.2307/2845334>
- Kennedy, R.E., Townsend, P.A., Gross, J.E., Cohen, W.B., Bolstad, P., Wang, Y.Q., Adams, P., 2009. Remote sensing change detection tools for natural resource managers: Understanding concepts and tradeoffs in the design of landscape monitoring projects. *Remote Sens. Environ., Monitoring Protected Areas* 113, 1382–1396. <https://doi.org/10.1016/j.rse.2008.07.018>
- Koenig, K., Höfle, B., 2016. Full-Waveform Airborne Laser Scanning in Vegetation Studies—A Review of Point Cloud and Waveform Features for Tree Species Classification. *Forests* 7, 198. <https://doi.org/10.3390/f7090198>
- Lallias-Tacon, S., Liébault, F., Piégay, H., 2017. Use of airborne LiDAR and historical aerial photos for characterising the history of braided river floodplain morphology and vegetation responses. *CATENA, Geoecology in Mediterranean mountain areas. Tribute to Professor José María García Ruiz* 149, 742–759. <https://doi.org/10.1016/j.catena.2016.07.038>
- Landon, N., Piégay, H., Bravard, J.P., 1998. The Drôme river incision (France): from assessment to management. *Landsc. Urban Plan.* 43, 119–131. [https://doi.org/10.1016/S0169-2046\(98\)00046-2](https://doi.org/10.1016/S0169-2046(98)00046-2)
- Laslier, M., Hubert-Moy, L., Dufour, S., 2019. Mapping Riparian Vegetation Functions Using 3D Bispectral LiDAR Data. *Water* 11, 483. <https://doi.org/10.3390/w11030483>
- Lee, P., Smyth, C., Boutin, S., 2004. Quantitative review of riparian buffer width guidelines from Canada and the United States. *J. Environ. Manage.* 70, 165–180. <https://doi.org/10.1016/j.jenvman.2003.11.009>
- Loicq, P., Moatar, F., Jullian, Y., Dugdale, S.J., Hannah, D.M., 2018. Improving representation of riparian vegetation shading in a regional stream temperature model using LiDAR data. *Sci. Total Environ.* 624, 480–490. <https://doi.org/10.1016/j.scitotenv.2017.12.129>
- Macfarlane, W.W., McGinty, C.M., Laub, B.G., Gifford, S.J., 2017. High-resolution riparian vegetation mapping to prioritize conservation and restoration in an impaired desert river. *Restor. Ecol.* 25, 333–341. <https://doi.org/10.1111/rec.12425>
- Maillard, P., Alencar-Silva, T., 2013. A method for delineating riparian forests using region-based image classification and depth-to-water analysis. *Int. J. Remote Sens.* 34, 7991–8010. <https://doi.org/10.1080/01431161.2013.827847>
- Manfreda, S., McCabe, M.F., Miller, P.E., Lucas, R., Pajuelo Madrigal, V., Mallinis, G., Ben Dor, E., Helman, D., Estes, L., Ciraolo, G., Müllerová, J., Tauro, F., De Lima, M.I., De Lima, J.L.M.P., Maltese, A., Frances, F., Caylor, K., Kohv, M., Perks, M., Ruiz-Pérez, G., Su, Z., Vico, G., Toth, B., 2018. On the Use of Unmanned Aerial Systems for Environmental Monitoring. *Remote Sens.* 10, 641. <https://doi.org/10.3390/rs10040641>
- Manners, R., Schmidt, J., Wheaton, J.M., 2013. Multiscalar model for the determination of spatially explicit riparian vegetation roughness. *J. Geophys. Res. Earth Surf.* 118, 65–83. <https://doi.org/10.1029/2011JF002188>
- Marcus, W.A., Fonstad, M.A., 2008. Optical remote mapping of rivers at sub-meter resolutions and watershed extents. *Earth Surf. Process. Landf.* 33, 4–24. <https://doi.org/10.1002/esp.1637>
- Martin, F.-M., Müllerová, J., Borgniet, L., Dommange, F., Breton, V., Evette, A., 2018. Using Single- and Multi-Date UAV and Satellite Imagery to Accurately Monitor Invasive Knotweed Species. *Remote Sens.* 10, 1662. <https://doi.org/10.3390/rs10101662>
- Matsuura, T., Suzuki, W., 2013. Analysis of topography and vegetation distribution using a digital elevation model: case study of a snowy mountain basin in northeastern Japan. *Landsc. Ecol. Eng.* 9, 143–155. <https://doi.org/10.1007/s11355-012-0187-2>
- Miao, X., Patil, R., Heaton, J.S., Tracy, R.C., 2011. Detection and classification of invasive saltcedar through high spatial resolution airborne hyperspectral imagery. *Int. J. Remote Sens.* 32, 2131–2150. <https://doi.org/10.1080/01431161003674618>
- Michez, A., Piégay, H., Jonathan, L., Claessens, H., Lejeune, P., 2016a. Mapping of riparian invasive species with supervised classification of Unmanned Aerial System (UAS) imagery. *Int. J. Appl. Earth Obs. Geoinformation* 44, 88–94. <https://doi.org/10.1016/j.jag.2015.06.014>
- Michez, A., Piégay, H., Lejeune, P., Claessens, H., 2017a. Multi-temporal monitoring of a regional riparian buffer network (>12,000 km) with LiDAR and photogrammetric point clouds. *J. Environ. Manage.* 202, 424–436. <https://doi.org/10.1016/j.jenvman.2017.02.034>

- Michez, A., Piégay, H., Lejeune, P., Claessens, H., 2017b. Multi-temporal monitoring of a regional riparian buffer network (>12,000 km) with LiDAR and photogrammetric point clouds. *J. Environ. Manage.*, Piégay & Lamouroux “Enlarging spatial and temporal scales for biophysical diagnosis and sustainable river management” 202, 424–436. <https://doi.org/10.1016/j.jenvman.2017.02.034>
- Michez, A., Piégay, H., Lisein, J., Claessens, H., Lejeune, P., 2016b. Classification of riparian forest species and health condition using multi-temporal and hyperspatial imagery from unmanned aerial system. *Environ. Monit. Assess.* 188, 146. <https://doi.org/10.1007/s10661-015-4996-2>
- Morgan, J.L., Gergel, S.E., Coops, N.C., 2010. Aerial Photography: A Rapidly Evolving Tool for Ecological Management. *BioScience* 60, 47–59. <https://doi.org/10.1525/bio.2010.60.1.9>
- Muller, E., Décamps, H., Dobson, M.K., 1993. Contribution of space remote sensing to river studies. *Freshw. Biol.* 29, 301–312. <https://doi.org/10.1111/j.1365-2427.1993.tb00766.x>
- Munné, A., Prat, N., Solà, C., Bonada, N., Rieradevall, M., 2003. A simple field method for assessing the ecological quality of riparian habitat in rivers and streams: QBR index. *Aquat. Conserv. Mar. Freshw. Ecosyst.* 13, 147–163. <https://doi.org/10.1002/aqc.529>
- Nagler, P.L., Brown, T., Hultine, K.R., van Riper, C., Bean, D.W., Dennison, P.E., Murray, R.S., Glenn, E.P., 2012. Regional scale impacts of Tamarix leaf beetles (*Diorhabda carinulata*) on the water availability of western U.S. rivers as determined by multi-scale remote sensing methods. *Remote Sens. Environ.* 118, 227–240. <https://doi.org/10.1016/j.rse.2011.11.011>
- Nagler, P.L., Doody, T.M., Glenn, E.P., Jarchow, C.J., Barreto-Muñoz, A., Didan, K., 2016. Wide-area estimates of evapotranspiration by red gum (*Eucalyptus camaldulensis*) and associated vegetation in the Murray–Darling River Basin, Australia. *Hydrol. Process.* 30, 1376–1387. <https://doi.org/10.1002/hyp.10734>
- Naiman, R.J., Décamps, H., 1997. The Ecology of Interfaces: Riparian Zones. *Annu. Rev. Ecol. Syst.* 28, 621–658. <https://doi.org/10.1146/annurev.ecolsys.28.1.621>
- Naiman, R.J., Decamps, H., McClain, M.E., 2005. *Riparia: Ecology, Conservation, and Management of Streamside Communities*. Elsevier.
- Nardi, F., Annis, A., Di Baldassarre, G., Vivoni, E.R., Grimaldi, S., 2019. GFPLAIN250m, a global high-resolution dataset of Earth’s floodplains. *Sci. Data* 6, 180309. <https://doi.org/10.1038/sdata.2018.309>
- Narumalani, S., Mishra, D.R., Wilson, R., Reece, P., Kohler, A., 2009. Detecting and Mapping Four Invasive Species along the Floodplain of North Platte River, Nebraska. *Weed Technol.* 23, 99–107. <https://doi.org/10.1614/WT-08-007.1>
- Palmquist, E.C., Ralston, B.E., Merritt, D.M., Shafroth, P.B., 2018. Landscape-scale processes influence riparian plant composition along a regulated river. *J. Arid Environ.* 148, 54–64. <https://doi.org/10.1016/j.jaridenv.2017.10.001>
- Parent, J.R., Volin, J.C., Civco, D.L., 2015. A fully-automated approach to land cover mapping with airborne LiDAR and high resolution multispectral imagery in a forested suburban landscape. *ISPRS J. Photogramm. Remote Sens.* 104, 18–29. <https://doi.org/10.1016/j.isprsjprs.2015.02.012>
- Peerbhay, K., Mutanga, O., Lottering, R., Ismail, R., 2016. Mapping *Solanum mauritianum* plant invasions using WorldView-2 imagery and unsupervised random forests. *Remote Sens. Environ.* 182, 39–48. <https://doi.org/10.1016/j.rse.2016.04.025>
- Penning, E., 2018. Interactions between flow and vegetation: Translating knowledge from academic research to daily water management. *E3S Web Conf.* 40, 01001. <https://doi.org/10.1051/e3sconf/20184001001>
- Piégay, H., Landon, N., 1997. Promoting ecological management of riparian forests on the Drôme River, France. *Aquat. Conserv. Mar. Freshw. Ecosyst.* 7, 287–304. [https://doi.org/10.1002/\(SICI\)1099-0755\(199712\)7:4<287::AID-AQC247>3.0.CO;2-S](https://doi.org/10.1002/(SICI)1099-0755(199712)7:4<287::AID-AQC247>3.0.CO;2-S)
- Pontius, R.G.J., Millones, M., 2011. Death to Kappa: birth of quantity disagreement and allocation disagreement for accuracy assessment. *Int. J. Remote Sens.* 32, 4407–4429. <https://doi.org/10.1080/01431161.2011.552923>
- Poole, G.C., Berman, C.H., 2001. An Ecological Perspective on In-Stream Temperature: Natural Heat Dynamics and Mechanisms of Human-Caused Thermal Degradation. *Environ. Manage.* 27, 787–802. <https://doi.org/10.1007/s002670010188>
- Rajão, R., 2013. Representations and discourses: the role of local accounts and remote sensing in the formulation of Amazonia’s environmental policy. *Environ. Sci. Policy, SI: Environmental and Developmental Discourses: Technical knowledge, discursive spaces and politics* 30, 60–71. <https://doi.org/10.1016/j.envsci.2012.07.019>
- Rapinel, S., Mony, C., Lecoq, L., Clément, B., Thomas, A., Hubert-Moy, L., 2019. Evaluation of Sentinel-2 time-series for mapping floodplain grassland plant communities. *Remote Sens. Environ.* 223, 115–129. <https://doi.org/10.1016/j.rse.2019.01.018>

- Richardson, D.M., Holmes, P.M., Esler, K.J., Galatowitsch, S.M., Stromberg, J.C., Kirkman, S.P., Pyšek, P., Hobbs, R.J., 2007. Riparian vegetation: degradation, alien plant invasions, and restoration prospects. *Divers. Distrib.* 13, 126–139. <https://doi.org/10.1111/j.1366-9516.2006.00314.x>
- Richardson, J.J., Moskal, L.M., Kim, S.-H., 2009. Modeling approaches to estimate effective leaf area index from aerial discrete-return LIDAR. *Agric. For. Meteorol.* 149, 1152–1160. <https://doi.org/10.1016/j.agrformet.2009.02.007>
- Richardson, J.J., Torgersen, C.E., Moskal, L.M., 2019. Lidar-based approaches for estimating solar insolation in heavily forested streams. *Hydrol. Earth Syst. Sci.* 23, 2813–2822. <https://doi.org/10.5194/hess-23-2813-2019>
- Richter, R., Reu, B., Wirth, C., Doktor, D., Vohland, M., 2016. The use of airborne hyperspectral data for tree species classification in a species-rich Central European forest area. *Int. J. Appl. Earth Obs. Geoinformation* 52, 464–474. <https://doi.org/10.1016/j.jag.2016.07.018>
- Riedler, B., Pernkopf, L., Strasser, T., Lang, S., Smith, G., 2015. A composite indicator for assessing habitat quality of riparian forests derived from Earth observation data. *Int. J. Appl. Earth Obs. Geoinformation, Special Issue on Earth observation for habitat mapping and biodiversity monitoring* 37, 114–123. <https://doi.org/10.1016/j.jag.2014.09.006>
- Rijkswaterstaat, 2014. Vegetatielegger : instrument voor veilige en natuurlijke uiterwaarden.
- Rood, S.B., Braatne, J.H., Goater, L.A., 2010. Responses of obligate versus facultative riparian shrubs following river damming. *River Res. Appl.* 26, 102–117. <https://doi.org/10.1002/rra.1246>
- Rubol, S., Ling, B., Battiato, I., 2018. Universal scaling-law for flow resistance over canopies with complex morphology. *Sci. Rep.* 8, 1–15. <https://doi.org/10.1038/s41598-018-22346-1>
- Rutherford, J.C., Meleason, M.A., Davies-Colley, R.J., 2018. Modelling stream shade: 2. Predicting the effects of canopy shape and changes over time. *Ecol. Eng.* 120, 487–496. <https://doi.org/10.1016/j.ecoleng.2018.07.008>
- Sá, C., Grieco, J., 2016. Open Data for Science, Policy, and the Public Good. *Rev. Policy Res.* 33, 526–543. <https://doi.org/10.1111/ropr.12188>
- Sankey, T.T., Sankey, J.B., Horne, R., Bedford, A., 2016. Remote Sensing of Tamarisk Biomass, Insect Herbivory, and Defoliation: Novel Methods in the Grand Canyon Region, Arizona [WWW Document]. <https://doi.org/info:doi/10.14358/PERS.82.8.645>
- Scott, M.L., Nagler, P.L., Glenn, E.P., Valdes-Casillas, C., Erker, J.A., Reynolds, E.W., Shafroth, P.B., Gomez-Limon, E., Jones, C.L., 2009. Assessing the extent and diversity of riparian ecosystems in Sonora, Mexico. *Biodivers. Conserv.* 18, 247–269. <https://doi.org/10.1007/s10531-008-9473-6>
- Seddon, J.A., Zenger, A., Doyle, S.J., Briggs, S.V., 2007. The extent of dryland salinity in remnant woodland and forest within an agricultural landscape. *Aust. J. Bot.* 55, 533–540. <https://doi.org/10.1071/BT06100>
- Shendryk, I., Broich, M., Tulbure, M.G., McGrath, A., Keith, D., Alexandrov, S.V., 2016. Mapping individual tree health using full-waveform airborne laser scans and imaging spectroscopy: A case study for a floodplain eucalypt forest. *Remote Sens. Environ.* 187, 202–217. <https://doi.org/10.1016/j.rse.2016.10.014>
- Shields, F.D., Coulton, K.G., Nepf, H., 2017. Representation of Vegetation in Two-Dimensional Hydrodynamic Models. *J. Hydraul. Eng.* 143, 02517002. [https://doi.org/10.1061/\(ASCE\)HY.1943-7900.0001320](https://doi.org/10.1061/(ASCE)HY.1943-7900.0001320)
- Sims, N.C., Colloff, M.J., 2012. Remote sensing of vegetation responses to flooding of a semi-arid floodplain: Implications for monitoring ecological effects of environmental flows. *Ecol. Indic.* 18, 387–391. <https://doi.org/10.1016/j.ecolind.2011.12.007>
- Solins, J.P., Thorne, J.H., Cadenasso, M.L., 2018. Riparian canopy expansion in an urban landscape: Multiple drivers of vegetation change along headwater streams near Sacramento, California. *Landsc. Urban Plan.* 172, 37–46. <https://doi.org/10.1016/j.landurbplan.2017.12.005>
- Sridhar, B.B.M., Vincent, R.K., Clapham, W.B., Sritharan, S.I., Osterberg, J., Neale, C.M.U., Watts, D.R., 2010. Mapping saltcedar (*Tamarix ramosissima*) and other riparian and agricultural vegetation in the Lower Colorado River region using multi-spectral Landsat TM imagery. *Geocarto Int.* 25, 649–662. <https://doi.org/10.1080/10106049.2010.521857>
- Staben, G.W., Evans, K.G., 2008. Estimates of tree canopy loss as a result of Cyclone Monica, in the Magela Creek catchment northern Australia. *Austral Ecol.* 33, 562–569. <https://doi.org/10.1111/j.1442-9993.2008.01911.x>
- Steiniger, S., Hay, G.J., 2009. Free and open source geographic information tools for landscape ecology. *Ecol. Inform.* 4, 183–195. <https://doi.org/10.1016/j.ecoinf.2009.07.004>

- Stella, J.C., Bendix, J., 2019. Chapter 5 - Multiple Stressors in Riparian Ecosystems, in: Sabater, S., Elosegí, A., Ludwig, R. (Eds.), *Multiple Stressors in River Ecosystems*. Elsevier, pp. 81–110. <https://doi.org/10.1016/B978-0-12-811713-2.00005-4>
- Stöcker, C., Bennett, R., Nex, F., Gerke, M., Zevenbergen, J., 2017. Review of the Current State of UAV Regulations. *Remote Sens.* 9, 459. <https://doi.org/10.3390/rs9050459>
- Straatsma, M.W., Baptist, M.J., 2008. Floodplain roughness parameterization using airborne laser scanning and spectral remote sensing. *Remote Sens. Environ.* 112, 1062–1080. <https://doi.org/10.1016/j.rse.2007.07.012>
- Straatsma, M.W., Fliervoet, J.M., Kabout, J.A.H., Baart, F., Kleinhans, M.G., 2019. Towards multi-objective optimization of large-scale fluvial landscaping measures. *Nat. Hazards Earth Syst. Sci.* 19, 1167–1187. <https://doi.org/10.5194/nhess-19-1167-2019>
- Straatsma, M.W., Kleinhans, M.G., 2018. Flood hazard reduction from automatically applied landscaping measures in RiverScape, a Python package coupled to a two-dimensional flow model. *Environ. Model. Softw.* 101, 102–116. <https://doi.org/10.1016/j.envsoft.2017.12.010>
- Strasser, T., Lang, S., 2015. Object-based class modelling for multi-scale riparian forest habitat mapping. *Int. J. Appl. Earth Obs. Geoinformation* 37, 29–37.
- Tabacchi, E., Correll, D.L., Hauer, R., Pinay, G., Planty-Tabacchi, A.-M., Wissmar, R.C., 1998. Development, maintenance and role of riparian vegetation in the river landscape. *Freshw. Biol.* 40, 497–516. <https://doi.org/10.1046/j.1365-2427.1998.00381.x>
- Tillack, A., Clasen, A., Kleinschmit, B., Förster, M., 2014. Estimation of the seasonal leaf area index in an alluvial forest using high-resolution satellite-based vegetation indices. *Remote Sens. Environ.* 141, 52–63. <https://doi.org/10.1016/j.rse.2013.10.018>
- Tompalski, P., Coops, N.C., White, J.C., Wulder, M.A., Yuill, A., 2017. Characterizing streams and riparian areas with airborne laser scanning data. *Remote Sens. Environ.* 192, 73–86. <https://doi.org/10.1016/j.rse.2017.01.038>
- Tomsett, C., Leyland, J., 2019. Remote sensing of river corridors: A review of current trends and future directions. *River Res. Appl.* 35, 779–803. <https://doi.org/10.1002/rra.3479>
- Tormos, T., Kosuth, P., Durrieu, S., Villeneuve, B., Wasson, J.G., 2011. Improving the quantification of land cover pressure on stream ecological status at the riparian scale using High Spatial Resolution Imagery. *Phys. Chem. Earth Parts ABC* 36, 549–559.
- Townsend, P.A., 2002. Relationships between forest structure and the detection of flood inundation in forested wetlands using C-band SAR. *Int. J. Remote Sens.* 23, 443–460. <https://doi.org/10.1080/01431160010014738>
- Townsend, P.A., Walsh, S.J., 2001. Remote sensing of forested wetlands: application of multitemporal and multispectral satellite imagery to determine plant community composition and structure in southeastern USA. *Plant Ecol.* 157, 129–149. <https://doi.org/10.1023/A:1013999513172>
- Turner, M.D., Taylor, P.J., 2003. Critical Reflections on the Use of Remote Sensing and GIS Technologies in Human Ecological Research. *Hum. Ecol.* 31, 177–182. <https://doi.org/10.1023/A:1023958712140>
- Turner, W., Rondinini, C., Pettorelli, N., Mora, B., Leidner, A.K., Szantoi, Z., Buchanan, G., Dech, S., Dwyer, J., Herold, M., Koh, L.P., Leimgruber, P., Taubenboeck, H., Wegmann, M., Wikelski, M., Woodcock, C., 2015. Free and open-access satellite data are key to biodiversity conservation. *Biol. Conserv.* 182, 173–176. <https://doi.org/10.1016/j.biocon.2014.11.048>
- Vande Kamp, K., Rigge, M., Troelstrup, N.H., Smart, A.J., Wylie, B., 2013. Detecting Channel Riparian Vegetation Response to Best-Management-Practices Implementation in Ephemeral Streams With the Use of Spot High-Resolution Visible Imagery. *Rangel. Ecol. Manag.* 66, 63–70. <https://doi.org/10.2111/REM-D-11-00153.1>
- Vanden Borre, J., Paelinckx, D., Múcher, C.A., Kooistra, L., Haest, B., De Blust, G., Schmidt, A.M., 2011. Integrating remote sensing in Natura 2000 habitat monitoring: Prospects on the way forward. *J. Nat. Conserv.* 19, 116–125. <https://doi.org/10.1016/j.jnc.2010.07.003>
- Varga, K., Dévai, G., Tóthmérész, B., 2013. Land use history of a floodplain area during the last 200 years in the Upper-Tisza region (Hungary). *Reg. Environ. Change* 13, 1109–1118. <https://doi.org/10.1007/s10113-013-0424-8>
- Vaz, A.S., Alcaraz-Segura, D., Campos, J.C., Vicente, J.R., Honrado, J.P., 2018. Managing plant invasions through the lens of remote sensing: A review of progress and the way forward. *Sci. Total Environ.* 642, 1328–1339. <https://doi.org/10.1016/j.scitotenv.2018.06.134>

- Verry, E.S., Dolloff, C.A., Manning, M.E., 2004. Riparian ecotone: A functional definition and delineation for resource assessment. *Water Air Soil Pollut. Focus* 4, 67–94. <https://doi.org/10.1023/B:WAFO.0000012825.77300.08>
- Vivier, A., Breton, L., Grivel, S., Melun, G., Piégay, H., Demarchi, L., Adrien, M., Séjour, A., Thommeret, N., Tormos, T., Cazals, C., Koehl, M., Lague, D., Rapinel, S., Ville, A., Chevillier, B., Tarrío, D., Guerri, O., Laslier, M., Dupont, P., 2018. Actes de la journée technique "Avancées, apports et perspectives de la télédétection pour la caractérisation physique des corridors fluviaux".
- Wagner-Lücker, I., Lanz, E., Förster, M., Janauer, G.A., Reiter, K., 2013. Knowledge-based framework for delineation and classification of ephemeral plant communities in riverine landscapes to support EC Habitat Directive assessment. *Ecol. Inform.*, The analysis and application of spatial ecological data to support the conservation of biodiversity 14, 44–47. <https://doi.org/10.1016/j.ecoinf.2012.11.003>
- Wallace, C.S.A., Villarreal, M.L., Iii, C. van R., 2013. Influence of monsoon-related riparian phenology on yellow-billed cuckoo habitat selection in Arizona. *J. Biogeogr.* 40, 2094–2107. <https://doi.org/10.1111/jbi.12167>
- Wan, Y., Wan, C., Hedgepeth, M., 2015. Elucidating multidecadal saltwater intrusion and vegetation dynamics in a coastal floodplain with artificial neural networks and aerial photography. *Ecology* 8, 309–324. <https://doi.org/10.1002/eco.1509>
- Wang, L., Silván-Cárdenas, J.L., Yang, J., Frazier, A.E., 2013. Invasive Saltcedar (*Tamarisk* spp.) Distribution Mapping Using Multiresolution Remote Sensing Imagery. *Prof. Geogr.* 65, 1–15. <https://doi.org/10.1080/00330124.2012.679440>
- Wasser, L., Chasmer, L., Day, R., Taylor, A., 2015. Quantifying land use effects on forested riparian buffer vegetation structure using LiDAR data. *Ecosphere* 6, art10. <https://doi.org/10.1890/ES14-00204.1>
- Wawrzyniak, V., Allemand, P., Bailly, S., Lejot, J., Piégay, H., 2017. Coupling LiDAR and thermal imagery to model the effects of riparian vegetation shade and groundwater inputs on summer river temperature. *Sci. Total Environ.* 592, 616–626. <https://doi.org/10.1016/j.scitotenv.2017.03.019>
- Weissteiner, C.J., Ickerott, M., Ott, H., Probeck, M., Ramminger, G., Clerici, N., Dufourmont, H., De Sousa, A.M.R., 2016. Europe's Green Arteries—A Continental Dataset of Riparian Zones. *Remote Sens.* 8, 925. <https://doi.org/10.3390/rs8110925>
- Wohl, E., 2017. Bridging the gaps: An overview of wood across time and space in diverse rivers. *Geomorphology, Dynamics and ecology of Wood in World Rivers* 279, 3–26. <https://doi.org/10.1016/j.geomorph.2016.04.014>
- Wohl, E., Bledsoe, B.P., Fausch, K.D., Kramer, N., Bestgen, K.R., Gooseff, M.N., 2016. Management of Large Wood in Streams: An Overview and Proposed Framework for Hazard Evaluation. *JAWRA J. Am. Water Resour. Assoc.* 52, 315–335. <https://doi.org/10.1111/1752-1688.12388>
- Wulder, M.A., White, J.C., Nelson, R.F., Næsset, E., Ørka, H.O., Coops, N.C., Hilker, T., Bater, C.W., Gobakken, T., 2012. Lidar sampling for large-area forest characterization: A review. *Remote Sens. Environ.* 121, 196–209. <https://doi.org/10.1016/j.rse.2012.02.001>
- Yang, X., 2007. Integrated use of remote sensing and geographic information systems in riparian vegetation delineation and mapping. *Int. J. Remote Sens.* 28, 353–370. <https://doi.org/10.1080/01431160600726763>
- Zahidi, I., Yusuf, B., Cope, M., Ahmed Mohamed, T., Mohd Shafri, H.Z., 2018. Effects of depth-varying vegetation roughness in two-dimensional hydrodynamic modelling. *Int. J. River Basin Manag.* 16, 413–426. <https://doi.org/10.1080/15715124.2017.1394313>
- Zogaris, S., Markogianni, V., Özeren, S.C., Dimitriou, E., 2015. Assessment of riparian zone and river island conditions in a trans-boundary greenbelt: the Evris/Meriç River (Greece - Turkey). *Fresenius Environ. Bull.* 24, 10.